

1                    **Attributing urban evapotranspiration from**  
2                    **eddy-covariance to surface cover: bottom-up versus**  
3                    **top-down**

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18                    **Key Points:**

- 19                    • Urban eddy-covariance footprints distinctly differ in surface cover composition from  
20                    hour to hour.
- 21                    • Impervious surfaces evaporate less than their surface fraction, but their contribu-  
22                    tion cannot be neglected.
- 23                    • High vegetation contributes up to 50% more to the total evaporation than its sur-  
24                    face fraction in this study.

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**Abstract**

Evapotranspiration ( $ET$ ) is a key process in the hydrological cycle that can help mitigate urban heat.  $ET$  depends on the surface cover, as the surface affects the partitioning of precipitation between runoff and evapotranspiration. In urban neighborhoods, this surface cover is highly heterogeneous. The resulting neighborhood-scale  $ET$  is observed with eddy-covariance systems. However, these observations represent the signal from wind- and stability-dependent footprints resulting in a continuously changing surface cover composition. This continuous change prevents quantitative analysis of the separate types. Here, we disentangle this neighborhood-scale  $ET$  at two urban sites in Berlin attributing the  $ET$  dynamics to the four major surface cover types in the footprint: impervious surfaces, low vegetation, high vegetation, and open water. Starting from the surface, we reconstruct  $ET$  based on patch-scale observations and conceptual models. Alternatively, we start with the eddy-covariance observations and attribute  $ET$  to the surface cover types solving a system of equations for four eddy-covariance systems with different footprints. Although starting at the surface yields more robust results, both approaches indicate that vegetation is responsible for more  $ET$  than proportional to its surface fraction, and evaporation from impervious surfaces although less cannot be neglected. We confirm the intuitive relation between  $ET$  and the surface cover fractions based on a wide range of surface compositions.

**Plain Language Summary**

Different types of surfaces, like grass, trees, pavement, and open water, affect how rainwater is divided between evaporation and runoff. In cities with lots of pavement and buildings, more water runs off than in natural areas leaving less water for evaporation. Measurement towers have been observing the evaporation from whole neighborhoods, but separating the effects of different surfaces is hard. In our study, we figure out how much each surface type contributes to evaporation with two methods: one starting from the separate surfaces and rebuilding the neighborhood evaporation, and the other starting with the neighborhood evaporation and breaking it down into evaporation from each surface. Both ways showed that plants evaporate more than proportionally to their surface area, but even built surfaces like pavement evaporate. Our findings confirm that more plants lead to more evaporation, but built surfaces cannot be ignored. This information can help urban planners create cities that manage water better, making cities nicer places to live.

**1 Introduction**

How precipitation is partitioned between runoff and evapotranspiration ( $ET$ ) plays an important role in the urban climate and is governed by the surface cover composition (Paul & Meyer, 2001; Oke et al., 2017). In cities, the abundant impervious surfaces prevent infiltration and promote surface runoff leaving less water available for  $ET$  than pervious areas (Fletcher et al., 2013; McGrane, 2016; Jongen et al., 2022). On the other hand, urban vegetation has the opposite effect increasing infiltration and  $ET$  (Peters et al., 2011; Gunawardena et al., 2017). While all vegetation favors  $ET$  compared to impervious surfaces, an isotope-based study revealed the vegetation type also affects infiltration and  $ET$  patterns (Kuhlemann et al., 2021). The combination of surface covers thus controls the water partitioning and consequently  $ET$  dynamics.

Promoting green surface covers by planting vegetation can increase  $ET$  using more of the available energy (Wang & Shu, 2020). Like vegetation, open water is suggested to potentially cool its surroundings by evaporation when implemented appropriately (Solcerova et al., 2019; Jacobs et al., 2020), although warming can also occur due to the high thermal inertia (Theeuwes et al., 2013; Steeneveld et al., 2014). The energy needed for the additional  $ET$  cannot heat the air mitigating heat and the associated health risks (Oke,

1982; Heaviside et al., 2017; Ward & Grimmond, 2017). However, how surface cover composition at the patch level ( $\sim 10^1 - 10^2$  m of a single surface cover type) translates to the neighborhood scale ( $\sim 10^3$  m) is largely unknown until now. To answer this question, we need to quantify how surface cover impacts the partitioning of incoming water fluxes (Bonneau et al., 2018) and how this affects the partitioning on the larger, neighborhood scale. Ultimately, the neighborhood scale is where the effect of the surface covers on *ET* needs to be understood. In time, this understanding will support the management of the cooling benefits and urban water demands.

At the neighborhood scale, eddy-covariance (EC) systems observe the *ET* of the combined surface cover types in their footprint (Feigenwinter et al., 2012). We refer to this as *ET*, since these observations show the combined signal of contributions from the present surface cover types and thus include evaporation, transpiration, and anthropogenic fluxes. Even though the heterogeneous urban surface results in spatially variable *ET* (Qin et al., 2022), the observed *ET* represents the weighted average flux in the footprint, as the EC systems are typically installed at a height where the heterogeneous surface flux sources are blended (Oke et al., 2017). Apart from this height, the footprint varies temporally depending on the wind speed and direction, and atmospheric stability (Kljun et al., 2015). Previous research demonstrated it is possible to upscale patch-scale *ET* observations to the neighborhood-scale EC observations weighed by surface cover in the footprint climatology at a relatively homogeneous urban site (Peters et al., 2011). However, hour-to-hour variation in the footprint contains useful information to understand *ET*. This time-dependent surface information has been applied to improve the model performance of urban *ET* machine learning models (Vulova et al., 2021). Thus, for the more common heterogeneous urban sites, the footprint is crucial information to disentangle the neighborhood-scale *ET* and attribute it to the different surface cover types.

EC footprints can be estimated with a variety of models. Large-eddy simulations (LES) or Lagrangian stochastic particle dispersion models (LPD) fully model the airflow to find the source area (LES: Leclerc et al. (1997); Wang and Davis (2008); LPD: Kljun et al. (2002); Hsieh et al. (2003); LES and LPD combined: Hellsten et al. (2015); Auvinen et al. (2017)). These models are both labor-intensive and computationally expensive, which limits their applicability to relatively short case studies (Vesala et al., 2008). To analyze longer time series, faster footprint models have been developed with an analytical approach relying on the surface-layer theory (e.g. Schuepp et al., 1990; Schmid & Oke, 1990; Kormann & Meixner, 2001). Their validity is restricted to certain turbulence intensities or stratifications. More recently, Kljun et al. (2015) developed a two-dimensional footprint parameterization that takes away these limitations. Their model yields robust results for most boundary layer conditions at any observation height within the surface layer. This model enables the identification of the flux's source area for a long time series with a wide range of atmospheric conditions. Therefore, this model is suitable to study the influence of the changing footprints on *ET*.

To study the influence of surface cover on *ET*, Peters et al. (2011) have described the seasonal patterns in urban *ET* from major plant-functional types (trees and turf grass). These two vegetation types explain the majority of *ET* variation. They also find that the surface fraction of a vegetation type is the most important factor determining its contribution to total *ET* underlining the importance of the EC footprint. They assume impervious surface evaporation can be neglected, while other studies show this assumption may not be valid (Ramamurthy & Bou-Zeid, 2014; Chen et al., 2023). Below, we will test the assumption by including evaporation from impervious surfaces. Moreover, while their analysis is focused on the seasonal timescale, we will consider the hourly timescale. The hourly *ET* dynamics play a key role in the urban climate experienced by urban citizens. As a verification, Peters et al. (2011) compared the sum of their *ET* components against EC observations, in essence reconstructing the *ET* signal from the bottom up.

127 Very few cities have observations of all relevant hydrometeorological states and fluxes  
128 across a range of surface covers. Berlin is a notable exception. In Berlin, meteorologi-  
129 cal observations are performed as part of two observatories: the Urban Climate Obser-  
130 vatory operated by the Chair of Climatology at the Technische Universität Berlin (<https://uco.berlin/en>,  
131 Scherer et al., 2019) and the Steglitz Urban Ecohydrological Observatory from the IGB  
132 Leibniz-Institute of Freshwater Ecology and Inland Fisheries (Kuhlemann et al., 2020,  
133 2021). Additionally, campaigns have added to this observation infrastructure, for exam-  
134 ple, with drone-based observations (Vulova et al., 2019) or with ground-based remote sens-  
135 ing (Zeeman et al., 2023). The elaborate observation infrastructure has resulted in nu-  
136 merous studies focusing on Berlin (e.g. Meier & Scherer, 2012; Fenner et al., 2014, 2023),  
137 of which many focused on  $ET$ . Kuhlemann et al. (2021) show based on soil isotopes that  
138  $ET$  differs depending on the vegetation type with more interception but less soil evap-  
139 oration for higher vegetation types. Subsequently, these isotope observations provide the  
140 means to evaluate modeled water partitioning quantifying yearly  $ET$  fluxes for different  
141 vegetation types (Gillefalk et al., 2021, 2022). Another modeling study applied a physics-  
142 based model to study hourly  $ET$  (Duarte Rocha et al., 2022), which after validation was  
143 combined with remotely-sensed vegetation characteristics to map  $ET$  for all of Berlin  
144 (Rocha et al., 2022). Vulova et al. (2021) achieved similar modeling skill with machine  
145 learning trained on meteorological and remote sensing data. Because of the research in-  
146 frastructure and the extensive literature, Berlin offers a unique setting to study the link  
147 between the surface cover and  $ET$ .

148 While the evaporation dynamics from various surface cover types have been investi-  
149 gated previously, few studies have addressed these issues across a range of surface cover  
150 types. These studies show that surface cover types have very different evaporation dy-  
151 namics. Four main surface cover types can be distinguished: impervious surface, low vegetation,  
152 high vegetation, and open water. Impervious surfaces only evaporate when wet directly  
153 after rainfall resulting in highly dynamic evaporation (Wouters et al., 2015). In contrast,  
154 vegetation can draw water from the soil sustaining  $ET$  long after rainfall (Teuling et al.,  
155 2006; Boese et al., 2019). Amongst vegetation, differences are seen with higher average  
156  $ET$  for higher vegetation with its higher leaf area density than for lower vegetation (Gillefalk  
157 et al., 2021). Sufficiently deep open water has more constant evaporation given the abun-  
158 dant water and high heat storage capacity that can provide energy in the absence of so-  
159 lar radiation (Jansen et al., 2022). The term  $ET$  is used for vegetation because the com-  
160 bined signal from transpiration, interception, and soil evaporation is considered. Over  
161 impervious and open water surfaces, only evaporation occurs so the term evaporation  
162 is used. We hypothesize these behaviors are combined at the neighborhood scale, as ob-  
163 served with EC, dependent on their relative contribution to the surface.

164 In this study, we aim to estimate the  $ET$  contribution of different surface cover types  
165 in the footprint profiting from the diverse observations in Berlin. With this, we will show  
166 how the footprint varies over time, how  $ET$  behaves for each surface cover type, the re-  
167 lation between the surface cover and neighborhood  $ET$ , and the contribution of each sur-  
168 face cover type to neighborhood  $ET$ . To study the contribution of each surface cover type  
169 to  $ET$ , we take both a bottom-up and a top-down approach to attribute the EC-observed  
170  $ET$  to the four dominant surface cover types. For the bottom-up approach, we recon-  
171 struct the EC signal by summing the estimated  $ET$  contribution of each surface cover  
172 type weighed by its contribution to the footprint. In this approach, the  $ET$  contribu-  
173 tion of each type is mimicked with conceptual models and small-scale observations. The  
174 top-down approach is based on a system of equations, in which each equation describes  
175 the surface cover composition of one EC system and its resulting flux. The resulting flux  
176 can be attributed to the surface cover types by solving the system of equations. We aim  
177 to reveal how the surface cover type influences neighborhood  $ET$  behavior and to indi-  
178 cate how altering surface cover may affect urban climate. Understanding the relation-  
179 ship between urban surface cover and  $ET$  can inform future climate-resilient urban de-  
180 sign.

## 2 Study sites

This study examines observations from the capital and largest city of Germany, Berlin, which has a population of 3.7 million spread over 891 km<sup>2</sup> (Amt für Statistik Berlin-Brandenburg, 2019). Situated in the east of Germany, the climate is temperate oceanic (Cfb) (Kottek et al., 2006). The closest weather station from the German Weather Service (DWD, Berlin-Tempelhof) recorded a long-term (1991-2020) mean annual rainfall of 585 mm and mean air temperature of 10.2 °C (DWD, 2021b). Here, we study the warm months (April until October) of the relatively dry year of 2019 with 492 mm of precipitation, in which an intense observation campaign was organized (Vulova et al., 2019). The warm months are studied as most *ET* occurs during this time.

Two sites in Berlin are studied here: a suburban one and one close to the city center. The first, suburban site is an urban research garden located in the southwest of the city at the Rothenburgstraße (ROTH, 52.457°N, 13.315°E, Figure 1a, (Vulova et al., 2021)). This site is an ICOS (Integrated Carbon Observation System) Associated Ecosystem Station (DE-BeR). Its surroundings within 1 km consist of 47% impervious surface, 19% low vegetation, 34% high vegetation, and no open water (see Sec. 3.1). At ROTH, a 40-meter tower holds three EC systems (IRGASON, Campbell Scientific) at 2, 30, and 40 meters. For all EC systems in this study, the resolution is 30 minutes. The observations are quality controlled according to the literature and only high-quality data (flag 0) is used (Foken et al., 2004). Additionally, sap flow was observed at six trees with FLGS-TDP XM1000 sap velocity logger systems (Dynamax Inc, Houston, USA), and soil moisture content was measured in two locations below high vegetation at three depths: 10-15, 40-50, and 90-100 cm (CS650 reflectometers, Campbell Scientific) (Kuhlemann et al., 2020). Finally, the leaf area index was measured monthly over three transects through high vegetation (LAI-2200, LI-COR, Lincoln, USA) (Vulova et al., 2019). Along each transect, leaf area index measurements were conducted at 1-meter intervals to capture the canopy variability, while walking in the same direction each time for standardization. A tripod on a balcony served as a reference for the above-canopy light conditions measuring every 10 seconds.

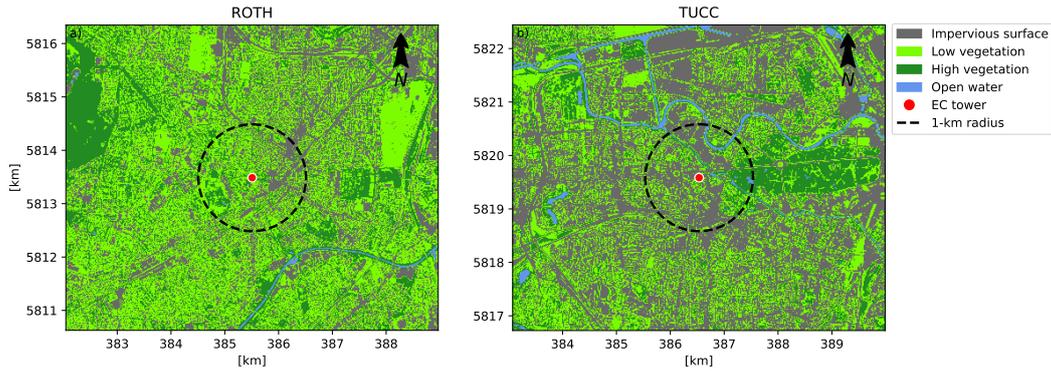
The second site is close to the city center at the TU Berlin Campus Charlottenburg (TUCC, 52.512°N, 13.328°E, Figure 1b, (Vulova et al., 2019; Jin et al., 2021)). Its surroundings within 1 km are more impervious than at ROTH: 62% impervious surface, 8% low vegetation, 26% high vegetation, and 3% open water (see Sec. 3.1). On the roof of TU Berlin (building height 46 meters), observations are made with a ceilometer (Lufft CHM 15k) and an EC system (IRGASON, Campbell Scientific). The EC system is attached to a 10-meter tower reaching 56 meters above ground level.

## 3 Methods

### 3.1 Surface cover classification

The surface cover needs to be classified to link the surface in the EC footprint to the neighborhood-scale *ET* observed with the EC system. Given the surface fraction covered by each surface cover type, the *ET* can be reconstructed from the evaporation dynamics of the different surface cover types (bottom-up, Figure 2a) or attributed to the surface cover types by linear decomposition (top-down, Figure 2b). For this study, we classify the surface into four different surface cover types: impervious surface, low vegetation, high vegetation, and open water. For this purpose, we combine information from four geospatial datasets from Berlin Open Data and the Berlin Digital Environmental Atlas:

- *Building height*: raster dataset at a 1-meter spatial resolution of all buildings in Berlin (Senate Department for Urban Development, Building and Housing, 2012).



**Figure 1.** Map of Berlin indicating the location of the (a) ROTH and (b) TUCC study sites with their surroundings classified in the four surface cover types distinguished in this study with the 1-km radius (dashed black line) around the EC towers (red dots). The coordinate reference system is WGS 84 UTM/33N EPSG: 32633)

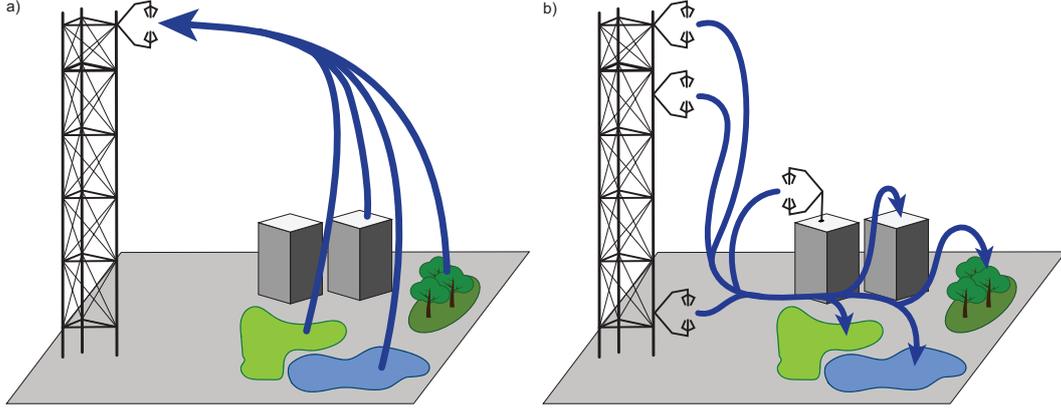
- 229 • *Vegetation height*: raster dataset at a 1-meter spatial resolution of all vegetation  
230 including trees, bushes, and grass in Berlin (Senate Department for Urban Develop-  
231 opment, Building and Housing, 2012)
- 232 • *Biotope types*: vector dataset describing the biotope type of all vegetation in Berlin  
233 according to the 7483 biotope types described by Zimmermann et al. (2015) (Senate  
234 Department for Urban Development, Building and Housing, 2013)
- 235 • *Streets*: vector dataset with all road segments in Berlin (Senate Department for  
236 the Environment, Mobility, Consumer and Climate Protection Berlin, 2014)

237 Around each EC tower, we classified the surface covers with a buffer of 0.025 de-  
238 grees latitude and 0.05 degrees longitude in both directions, equivalent to 2.8 and 3.4  
239 kilometers. In total, this gives an area of 5.6 by 6.8 kilometers. We selected this buffer,  
240 as for 90% of the footprints this area includes the entire footprint calculated in this study  
241 (see Sec. 3.2). For only 0.5% of the time, the buffer contains less than 80% of the foot-  
242 print. All datasets are clipped to this area. Vector datasets are resampled to rasters at  
243 a 1-meter resolution to ensure compatibility with the raster datasets.

244 At the start of the classification, all described vegetated land biotopes are assigned  
245 to vegetation and all water biotopes to open water. The impervious surface is determined  
246 based on all areas in the street data and all areas that have an assigned building height.  
247 The vegetation is split into low and high vegetation depending on the height with a thresh-  
248 old of 0.5 meters following Kuhleemann et al. (2021). The exact threshold has minimal  
249 influence as only a negligible fraction of the vegetation has a height between 0.3 and 1.0  
250 meters.

### 251 3.2 Footprint modeling

252 Footprints were calculated to determine the source area of the turbulent fluxes for  
253 all timesteps. We selected the flux footprint model from Kljun et al. (2015), which is fre-  
254 quently applied in urban environments (e.g. Stagakis et al., 2019; Nicolini et al., 2022;  
255 Karl et al., 2023). This footprint model provides two-dimensional grids with relative flux  
256 contribution. The model requires the measurement height, friction velocity, boundary-  
257 layer height, Obukhov length, wind direction, and mean and standard deviation of the  
258 wind speed. For all wind variables, EC observations are used, while the boundary-layer  
259 height is derived from ceilometer observations at the TUCC site. The Obukhov length



**Figure 2.** Conceptual drawing of the bottom-up (a) and top-down (b) approach. The arrows start at the data sources and end at the results of the approaches. Footprints determine the contribution for each surface cover type (not shown).

260 in  $m$  ( $L$ ) is calculated according to:

$$261 \quad L = -\frac{u_*^3 \bar{\theta}_v}{\kappa g (\overline{w'\theta'_v})_s} \quad (1)$$

262 where  $u_*$  the surface friction velocity in  $m\ s^{-1}$ ,  $\bar{\theta}_v$  the mean virtual potential temper-  
 263 ature in K,  $\kappa$  the von Kármán constant of 0.4,  $g$  the gravitational acceleration of  $9.81$   
 264  $m\ s^{-2}$ , and  $(\overline{w'\theta'_v})_s$  the kinematic virtual potential temperature flux in  $K\ m\ s^{-1}$  at the  
 265 observation height.

266 As the model results in contours per 10% and the 100%-contribution contour is in-  
 267 finite, the resulting footprint grids are limited to the 90%-contribution contour. Part of  
 268 the footprint is not taken into account when the footprint extends beyond the classified  
 269 area (Section 3.1). This last step had minimal influence, as the classified area is considerably  
 270 larger than the typically considered representative area within a radius of either 0.5 or  
 271 1 kilometer (Lipson et al., 2022). In the end, the surface fractions are calculated as the  
 272 footprint contribution per surface cover type taking into account the weight of each pixel.

### 273 3.3 Bottom-up

274 The bottom-up approach attributes  $ET$  to the different surface cover types by de-  
 275 termining evaporation dynamics for each type (Figure 2a). Consequently, the EC ob-  
 276 servations are hypothesized to be reconstructed when these dynamics are weighed with  
 277 the footprint contribution of the surface cover types. For the impervious surfaces, open  
 278 water, and high vegetation interception, evaporation dynamics are estimated based on  
 279 conceptual models. For the low vegetation and the high vegetation transpiration, obser-  
 280 vations capture the dynamics. We assume the evaporation dynamics per surface cover  
 281 type to be similar for ROTH and TUCC, as previous research found their forcing is com-  
 282 parable and can be used interchangeably to predict  $ET$  with the same accuracy (Duarte Rocha  
 283 et al., 2022). Negative  $ET$  observations are omitted, as the conceptual models are not  
 284 capable of predicting negative fluxes. This filter has a very limited impact on the results,

285 as it excludes only 384 of the 17780 30-minute time intervals. The results are analyzed  
 286 at two timescales, midday and daily, as these consider different aspects of  $ET$ . Midday  
 287 is defined from 11:00 until 15:00 local (10:00-14:00 UTC) time with every half hour con-  
 288 sidered separately. During these hours, incoming radiation driving  $ET$  is highest. Con-  
 289 sidering multiple hours minimizes the sampling noise due to the stochastic nature of tur-  
 290 bulance even at half-hourly timescales. The daily timescale is relevant for water resources  
 291 management.

### 292 *Impervious surface*

293 Evaporation from impervious surfaces is modeled according to Wouters et al. (2015).  
 294 Their parameterization includes two processes to mimic the water on an impervious sur-  
 295 face: rainfall and evaporation. The impervious surface is characterized by the maximum  
 296 water storage ( $w_m$ ) in  $\text{mm m}^{-2}$  and the maximum wet/evaporative fraction ( $\delta_m$ ). These  
 297 parameters were determined for Berlin based on 3D-LIDAR scans and found to be 1.03  
 298  $\text{mm m}^{-2}$  and 13.53% (Haacke, 2022). The evaporative fraction decreases following a power  
 299 law with an exponent of  $-\frac{2}{3}$  depending on the water storage, which follows from the as-  
 300 sumption that interception storage capacity linearly depends on the storage depth. Wa-  
 301 ter gain from rainfall is reduced in efficiency when closer to the water storage described  
 302 by:

$$303 \quad w(t + \Delta t) = w_m \left( 1 - \ln(1 - (1 - e^{(1 - \frac{w(t)}{w_m})})e^{-\frac{r_0 \Delta t}{w_m}}) \right) \quad (2)$$

304 where,  $w$  is the water storage in mm,  $t$  time in s,  $\Delta t$  length of the time step in s,  
 305 and  $r_0$  the rainfall intensity in  $\text{mm s}^{-1}$ . The formulation assumes constant rainfall dur-  
 306 ing a time step. The evaporation is described by:

$$307 \quad w(t + \Delta t) = \left( w(t)^{\frac{1}{3}} - \frac{\delta_m E_p \Delta t}{3w_m^{\frac{2}{3}}} \right)^3 \quad (3)$$

308 where  $E_p$  is the potential evaporation. The  $E_p$  is calculated according to Penman  
 309 (1956), further described in Eq. 4. The meteorological forcing has a resolution of 30 min-  
 310 utes, but the conceptual model is run numerically at a 30-second time step to ensure a  
 311 numerically robust solution with linearly interpolated meteorological forcing.

### 312 *High vegetation*

313 The  $ET$  from high vegetation consists of transpiration and interception. The tran-  
 314 spiration is derived from observations of the soil moisture content and sap flow as de-  
 315 scribed in Kuhlemann et al. (2021). Soil moisture content observations are used from both  
 316 the ‘‘Trees’’ and ‘‘Shrubs’’ plots for the transpiration estimation from high vegetation.  
 317 The soil moisture content reflects the evaporated water volume, but root water uptake  
 318 does not correlate directly with transpiration apparent from the lag between the two.  
 319 Therefore, we scale daily soil moisture loss with hourly sap flow observations. This method  
 320 takes advantage of the temporal variation in sap flow observations and the magnitude  
 321 of the soil moisture content observations. Soil moisture loss due to drainage is assumed  
 322 to be negligible, as the deepest soil moisture observations at 95 cm depth do not indi-  
 323 cate a drainage flux. Furthermore, soil moisture loss in the lowest layer of observations  
 324 is not added to the evaporation.

325 The canopy interception and its evaporation are modeled with the Rutter model  
 326 that allows for sub-daily resolution (Rutter et al., 1975; Valente et al., 1997). The model  
 327 partitions rainfall between evaporation from the canopy and trunk, throughfall, and stem  
 328 flow. Two storages are part of the model: the canopy and the trunk. Both storages evap-

329 orate at the potential rate calculated according to the Penman (1956) equation (Eq. 4).  
 330 Canopy storage capacity depends on the tree species ranging between 0.1 and 3 mm (e.g.  
 331 Aston, 1979; Klaassen et al., 1998; Baptista et al., 2018; Ramírez et al., 2018), although  
 332 in exceptional tropical canopies capacities up to 8 mm have been observed (Herwitz, 1985).  
 333 We assume the canopy storage capacity is linearly related to the leaf area index with a  
 334 storage of 0.2 mm per unit leaf area (Huang et al., 2017). Leaf area index observations  
 335 at monthly intervals are interpolated with a univariate spline with four degrees of free-  
 336 dom. The modeled interception appears to be relatively insensitive to the other para-  
 337 meters: trunk water storage capacity, partitioning between stem flow and throughfall,  
 338 and the fraction of evaporation from the stem flow. All of these parameters concern the  
 339 stem flow, which, on average, accounts for only 2% of the precipitation exceeding the canopy  
 340 storage capacity (Rutter et al., 1975). The modeled interception evaporation is added  
 341 to the transpiration to obtain the  $ET$  from the high vegetation.

### 342 ***Low vegetation***

343 Low vegetation is directly represented by an EC system installed at 2 meters di-  
 344 rectly above the grass at ROTH. In their similar study, Peters et al. (2011) installed an  
 345 EC system close to the surface to estimate the  $ET$  from low vegetation as well. Within  
 346 a forest, a comparable set-up helped to differentiate the  $ET$  components (Paul-Limoges  
 347 et al., 2020).

348 The quality-controlled  $ET$  is a direct observation of the low vegetation dynamics  
 349 when the wind comes from between east ( $90^\circ$ ) and southwest ( $230^\circ$ ). Fluxes were only  
 350 considered when suitable for process-focused studies (quality flag "0" according to Foken  
 351 et al. (2004)).

### 352 ***Open water***

353 Open water evaporation is estimated with a parameterization of the Penman (1956)  
 354 equation (De Bruin, 1979):

$$355 \quad E_p = 37 + 40\bar{u}_{2m}(e_{s,2m} - e_{40m}) \quad (4)$$

356 where  $\bar{u}_{2m}$  is the mean wind speed at 2 meters ( $\text{m s}^{-1}$ ),  $e_{s,2m}$  the saturated vapor pres-  
 357 sure at 2 meters (Pa), and  $e_{40m}$  the vapor pressure at 40 meters (Pa). Open water is as-  
 358 sumed to evaporate at the potential rate. In the case of a negative  $E_p$ , evaporation is  
 359 set to 0.

## 360 **3.4 Top-down**

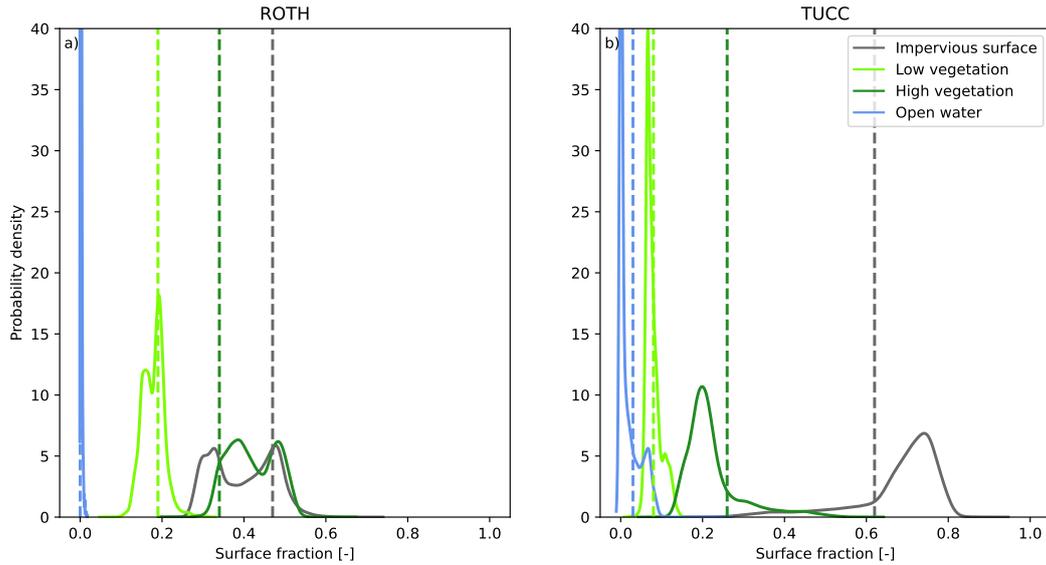
361 The top-down approach takes the neighborhood-scale EC observations and attributes  
 362 the flux to the different surface cover types by solving a system of equations (Figure 2b).  
 363 The system consists of four equations. Each equation describes how the surface covers  
 364 are combined according to the footprint to yield the EC observation. The evaporation  
 365 for the four surface cover types results in four unknowns, as the evaporation per surface  
 366 cover type is assumed similar for all EC systems. The linear system can be solved, as  
 367 it has an equal number of equations and unknowns.

$$368 \quad f_{im,1}E_{im} + f_{lv,1}E_{lv} + f_{hv,1}E_{hv} + f_{ow,1}E_{ow} = E_{EC,1} \quad (5)$$

$$369 \quad f_{im,2}E_{im} + f_{lv,2}E_{lv} + f_{hv,2}E_{hv} + f_{ow,2}E_{ow} = E_{EC,2} \quad (6)$$

$$370 \quad f_{im,3}E_{im} + f_{lv,3}E_{lv} + f_{hv,3}E_{hv} + f_{ow,3}E_{ow} = E_{EC,3} \quad (7)$$

$$371 \quad f_{im,4}E_{im} + f_{lv,4}E_{lv} + f_{hv,4}E_{hv} + f_{ow,4}E_{ow} = E_{EC,4} \quad (8)$$



**Figure 3.** Probability density of the time-dependent surface fractions in the EC footprint over the study period (April-October 2019). The dashed vertical lines indicate the average surface cover fraction within a 1-km radius of the EC (see Figure 1).

where  $f$  is the fraction of the impervious surface ( $im$ ), low vegetation ( $lv$ ), high vegetation ( $hv$ ), and open water ( $ow$ ), and  $E$  is evaporation of the same surfaces and the EC. The numbers indicate the different EC systems. The four EC systems are at the 56-m EC at TUCC and 2-, 30-, and 40-m ECs at ROTH. The evaporation from each surface can be determined given the fractions derived from the footprints and the EC observations. We exclude solutions with estimated evaporation below  $-3.5 \text{ mm d}^{-1}$  for one of the surface cover types, as these solutions likely have negative evaporation rates for one surface cover type that are balanced by positive evaporation rates for another type.

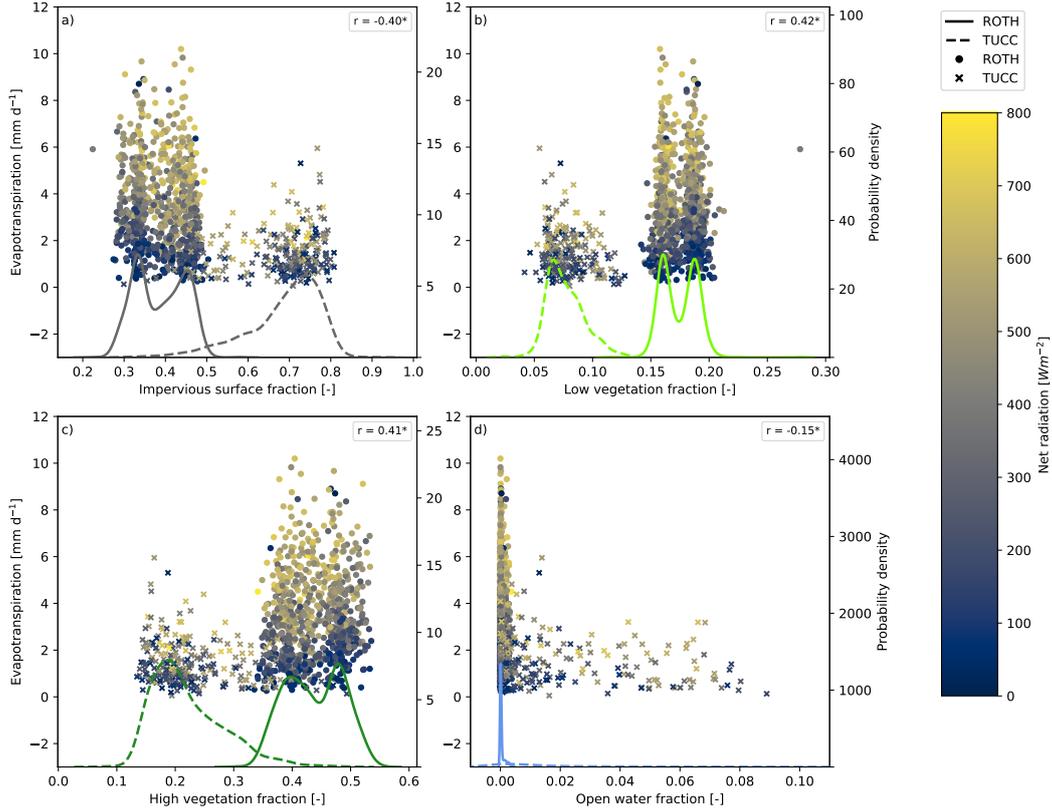
## 4 Results

### 4.1 Footprint variation

A high variation in footprint composition highlights the heterogeneity of the urban surface (Figure 3). The wide, non-normal distributions cause the actual surface fractions in the footprint to differ substantially from the surface cover fractions within a 1-km radius of the EC system (vertical lines) at most times. The 1-km radius estimation and the actual fraction are only similar for open water, as this covers a limited surface. For the impervious surface and high vegetation at ROTH, the bi-modality of the distribution demonstrates that a single value will not be able to capture the surface fractions. Additionally, surface covers can vary within a wide range as seen at TUCC where the impervious surface fraction varies from 0.2 up to 0.8. The high variation necessitates that the time-dependent footprint composition is considered to understand  $ET$  dynamics.

### 4.2 Surface cover composition impact on evapotranspiration

Combining the footprint variation from both sites with the neighborhood  $ET$  reveals the influence of the surface cover composition on  $ET$  (Figure 4). We find less  $ET$  with more impervious surface and more  $ET$  with more vegetation (high and low). Open water shows a less clear relation, as the open water fraction is very low most of the time.



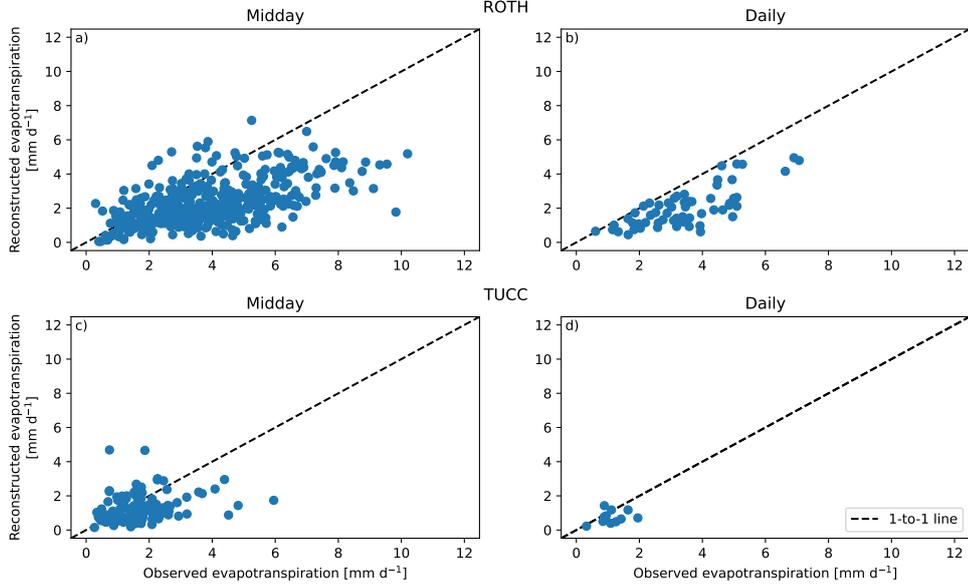
**Figure 4.** Relation of the observed  $ET$  and the surface cover fraction for each surface cover type (a: impervious surface, b: low vegetation, c: high vegetation, d: open water). The probability density curves (right axis) describe the footprint compositions for both ROTH (solid) and TUCC (dashed).

397 Although the surface cover is relevant, the variation in the  $ET$  indicates meteorologi-  
 398 cal conditions affect  $ET$  as well, illustrated by the ordering of the points by available en-  
 399 ergy quantified as the net radiation. While the surface cover composition in the footprint  
 400 varies at one site, the two sites together reveal an evident influence of the surface on  $ET$ .

### 401 4.3 Evapotranspiration attribution to the surface cover

402 Observed  $ET$  is approximated by  $ET$  reconstructed by a weighted average of sur-  
 403 face cover type evaporation dynamics (Figure 5 and Table 1). Performance depends on  
 404 surface cover, as results show consistently higher correlations at ROTH compared to the  
 405 more impervious TUCC. Additionally, the negative MBE indicates an underestimation  
 406 of total  $ET$ . The data gaps due to quality control of the 2-m EC system explain why  
 407 the number of evaluated data points is lower than the duration of the study period. In  
 408 two TUCC cases,  $ET$  is highly overestimated when a rainfall event coincides with a high  
 409 impervious fraction in the footprint and high potential evaporation (Figure 5c), for which  
 410 the conceptual model for impervious surfaces is responsible.

411 Absolute errors increase with rising  $ET$  rates across both timescales at ROTH (Fig-  
 412 ure 5). Other than with  $ET$  itself, the absolute errors correlate with the meteorologi-  
 413 cal and hydrological conditions (Table 2). The observed net radiation has a positive cor-  
 414 relation with the absolute  $ET$  error at ROTH while no correlation is present at TUCC.



**Figure 5.** Comparison of the *ET* observed with EC against the *ET* reconstructed with small-scale observations and conceptual models at (a-b) Rothenburgstraße and (c-d) TU Berlin campus for (a,c) midday hours per half hour and (b,d) daily means. Midday hours are between 11:00 and 15:00 local time (10:00-14:00 UTC). Table 1 gives an overview of the statistics.

**Table 1.** Overview of the performance of the bottom-up approach compared with EC *ET* observations per 30 minutes as shown in Figure 5.

		Rothenburgstraße		TU Berlin campus	
		<i>Midday</i>	<i>Daily</i>	<i>Midday</i>	<i>Daily</i>
Figure 5	panel	a	b	c	d
Data points	[-]	440	60	113	11
Observed mean <i>ET</i>	[mm d <sup>-1</sup> ]	3.9	3.4	1.7	1.1
Modeled mean <i>ET</i>	[mm d <sup>-1</sup> ]	2.3	2.0	1.3	0.8
Mean bias error	[mm d <sup>-1</sup> ]	-1.6	-1.4	-0.4	-0.4
Mean absolute error	[mm d <sup>-1</sup> ]	1.8	1.4	0.9	0.5
Pearson's r	[-]	0.56	0.76	0.29	0.27

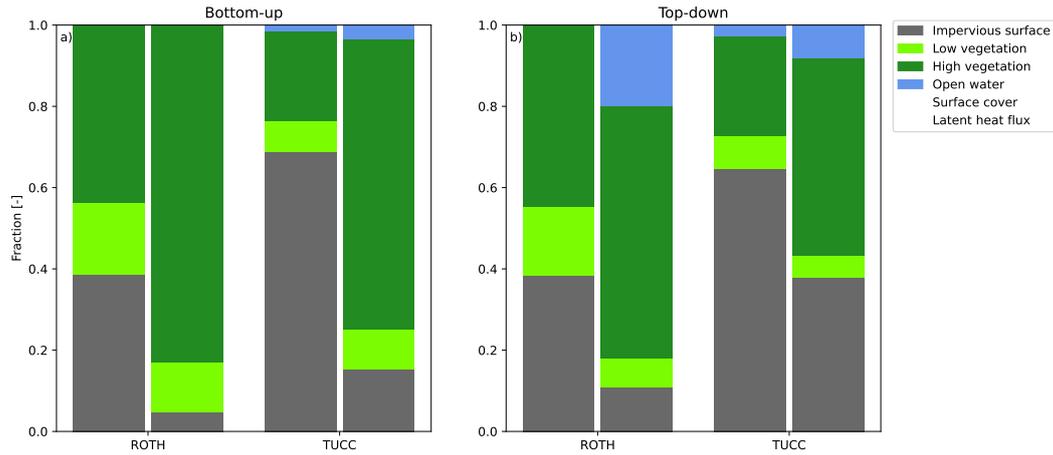
**Table 2.** Overview of Pearson correlations between environmental variables and the absolute error of the bottom-up reconstructed *ET*. Only significant correlations are shown (Wald’s test).

<b>Rothenburgstraße TU Berlin campus</b>				
	<i>Midday</i>	<i>Days</i>	<i>Midday</i>	<i>Days</i>
Net radiation	0.39	0.18		
Soil moisture	-0.34			
Specific humidity	0.35	0.20	0.16	
Vapour pressure deficit	0.45			
Impervious evaporation			0.59	0.86

415 This suggests the bottom-up approach performs worse with more available energy. At  
 416 the same time, soil moisture values show a negative correlation at ROTH suggesting er-  
 417 rors are smaller for relatively wet soil conditions. At TUCC, soil moisture does not cor-  
 418 relate with the error, which could be explained by the high impervious fraction. The spe-  
 419 cific humidity shows errors increase with more moist air. In contrast, vapor pressure deficit  
 420 indicates higher errors for a higher deficit of moisture (drier air) at ROTH. At TUCC,  
 421 high correlations are found with the evaporation from impervious surfaces indicating this  
 422 surface cover type might explain most of the errors. Overall, correlations are typically  
 423 weaker at TUCC except for evaporation from impervious surfaces. We expect the weaker  
 424 correlations to be a consequence of the lower MBE. In turn, the lower MBE may be par-  
 425 tially explained by the lower range of observed *ET* as a consequence of the high imper-  
 426 vious fraction.

427 Impervious surfaces contribute proportionally less to *ET* than their surface frac-  
 428 tion according to the bottom-up approach (Figure 6a). In contrast, high vegetation con-  
 429 tributes significantly more. The relative *ET* contribution of low vegetation varies depend-  
 430 ing on the composition of the remaining surfaces. In areas with mainly impervious sur-  
 431 faces, low vegetation exhibits a comparatively larger *ET* contribution, while in regions  
 432 mainly covered by high vegetation, its *ET* contribution is lower. Despite open water cov-  
 433 ering only a small fraction of the surface, the TUCC results indicate that the *ET* frac-  
 434 tion can exceed the surface fraction. The relative contributions are constant through-  
 435 out the months, although exact fraction values vary mostly around 0.02 with exceptions  
 436 up to 0.13. Throughout the study period, the surface fraction has the same qualitative  
 437 relation to *ET* contribution.

438 The top-down approach yields similar relative contributions to the surface cover  
 439 and *ET* as the bottom-up approach (Figure 6b). However, the *ET* fractions are more  
 440 similar to the surface fractions than the bottom-up approach indicates. Noteworthy, the  
 441 negligible open water surface contributes 20% to *ET* at ROTH. This result seems un-  
 442 likely given that nearly no water bodies are within the area of most footprints at ROTH.  
 443 Thus, it exposes a potential weakness of the top-down approach related to the data in-  
 444 stead of the physics-driven nature of this approach. On top of that, in only 44 timesteps  
 445 the linear system of equations resulted in a solution for two reasons. First, the data avail-  
 446 ability limits this approach to 342 timesteps. Subsequently, 298 timesteps are excluded  
 447 from the analysis as negative evaporation rates artificially enhanced the evaporation from  
 448 the other surfaces. This artificial enhancement is an artifact of the linear system of equa-  
 449 tions (Eq. 5). Next to the *ET* fractions, the surface fractions differ slightly from the bottom-  
 450 up approach because different timesteps are considered. Over the months, relative *ET*  
 451 contributions differ due to the low reliability of monthly estimates caused by the low data  
 452 availability (not shown). In some months, the relative *ET* contribution becomes nega-  
 453 tive, which we attribute to artificially-enhanced evaporation from the linear system. Un-  
 454 like the bottom-up approach, no direct comparison with observations can be made, as



**Figure 6.** Relative contribution of the surface cover types to total surface cover (vertical/horizontal hatch) and  $ET$  (diagonal hatch). Surface cover fractions differ between the two methods at the same site at different times and thus footprints are included in the analysis due to data availability.

455 the method gets the EC observations as input, and no observations are available at the  
 456 patch scale.

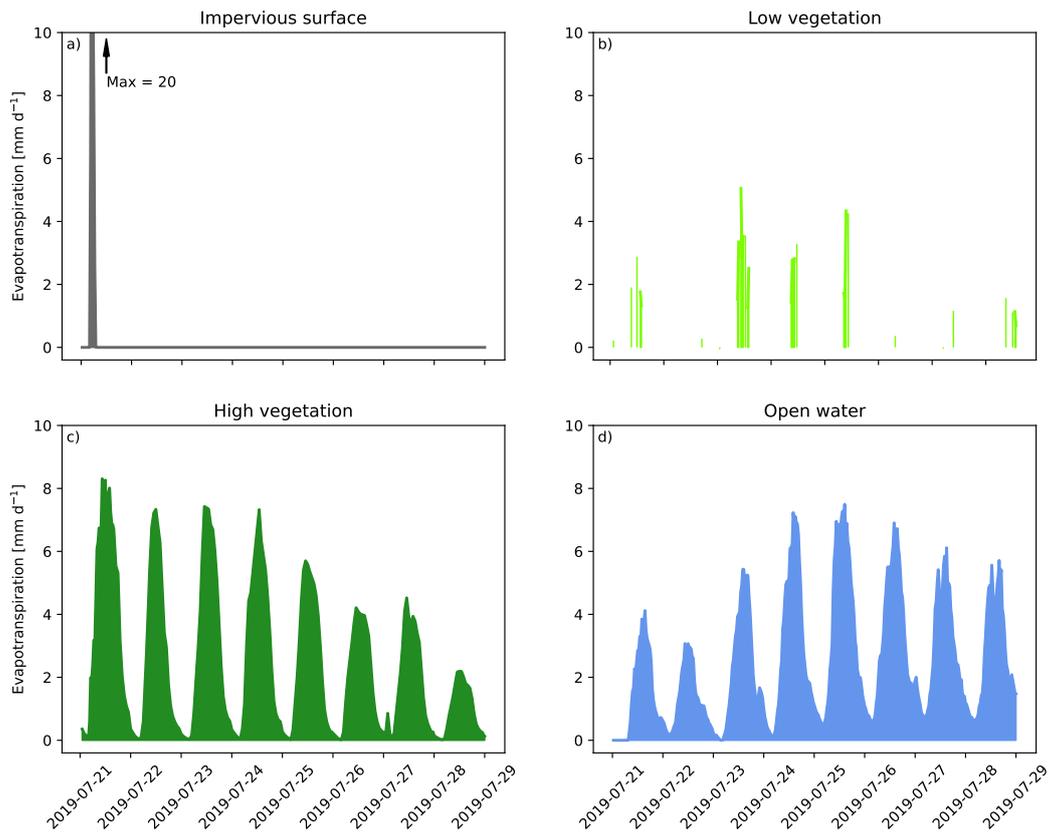
#### 457 4.4 Evaporation dynamics per surface cover type

458 The distinct evaporation dynamics of each surface cover type are visible when zooming  
 459 in on one drydown (Figure 7). These dynamics can be derived from the bottom-up  
 460 approach given its good performance and the high number of timesteps with attributed  
 461 fluxes. The impervious surface has a unique pattern with a sharp peak after rainfall and  
 462 no evaporation once the surface has dried. Meanwhile, the other three surface cover types  
 463 all show a daily cycle. Low vegetation and open water show comparable changes over  
 464 time without a response to the time since the last precipitation but following energy avail-  
 465 ability and transport efficiency. On the other hand, high vegetation limits  $ET$  within days  
 466 after rainfall. These responses are seen in all other drydowns except for the last drydown  
 467 during the warm season. At this time, the soil moisture is more depleted triggering low  
 468 vegetation to limit  $ET$ , while open water maintains the same response.

## 469 5 Discussion and conclusion

### 470 5.1 Surface cover type contributions to evapotranspiration

471 Our study revealed that the four distinguished surface cover types do not contribute  
 472 to  $ET$  proportional to their surface fraction. To disentangle these contributions, the  $ET$   
 473 was attributed to the surface covers with both a bottom-up and top-down approach. Both  
 474 approaches find similar  $ET$  contributions compared to the surface fraction; impervious  
 475 surfaces evaporate less than their surface fraction, while high vegetation and open wa-  
 476 ter evaporate more. For high vegetation, an isotope-evaluated model study found sim-  
 477 ilar ratios between surface fraction ( $\sim 30\%$ ) and  $ET$  contribution ( $\sim 80\%$ ) at ROTH (Gillefalk  
 478 et al., 2022). From this  $ET$ , evaporation of interception accounts for 17% of the total  
 479 precipitation over the study period from April to October. This is comparable to some  
 480 studies finding values between 14-27% (Bryant et al., 2005; Xiao & McPherson, 2011),  
 481 while others show higher interception evaporation between 45 and 77% (Asadian & Weiler,



**Figure 7.** Illustration of *ET* dynamics during a drydown starting 30 minutes after rainfall ceased determined with the bottom-up approach for the four surface cover types (a: impervious surface, b: low vegetation, c: high vegetation, d: open water). This drydown occurred between 21-07-2019 and 29-07-2019. The gaps in the are explained by the quality control of the 2-m EC measuring low vegetation.

2009; Anys & Weiler, 2023) or lower between 5-6% (Paul-Limoges et al., 2020). Although our interception evaporation is lower than most observed values, together with transpiration, it exceeds the precipitation during the study period. Soil moisture reserves supply the additional water. For low vegetation, the *ET* contribution may either be higher or lower than their surface fraction depending on the composition of the other surfaces.

The found *ET* contributions are largely in line with findings by Peters et al. (2011), who did a similar exercise for a more homogeneous neighborhood. Still, we challenge their assumption that the impervious surface did not contribute anything to *ET*, as we find 5% (11% top-down) of *ET* may come from impervious surfaces in a suburban setting (ROTH, 39% impervious in footprint). In the more impervious city center (TUCC, 65/69% impervious in the footprint), we find a contribution of 15% (38% top-down). Ramamurthy and Bou-Zeid (2014); Chen et al. (2023) found *ET* from impervious surfaces contributed between 11 and 18%.

Even though the *ET* contribution was similar for the bottom-up and the top-down approach, these methodologies also showed two interesting differences. Given these two differences, we think the bottom-up approach has the most potential to contribute to our understanding of the link between patch- and neighborhood-scale *ET*. The first difference is the number of timesteps with a successful outcome: 1112 for bottom-up and 44 for top-down. The maximum number is limited by the study period of 244 days equal to 11,712 half hours for top-down, of which 2,196 are during the midday hours. The EC observations cause the high number of timesteps without results, as these EC observations contain considerable gaps, as many observations are filtered during quality control because of the challenging urban environment (Feigenwinter et al., 2012; Oke et al., 2017). Moreover, only the highest-quality EC observations are suitable for our analysis (quality flag "0" (Foken et al., 2004)), as this study focuses on the process level. While data availability is a challenge for both approaches, the top-down approach relies more heavily on EC observations leading to even fewer timesteps with results.

The second difference is that the bottom-up approach is driven by physics, while the top-down approach is based on mathematics. The physics-driven bottom-up approach provides insight into the *ET* contributions of the surface cover types but still has a mismatch with the observed *ET*. Also using a bottom-up approach, Salmond et al. (2012) reconstructed the neighborhood-scale sensible heat flux observed with an EC system with smaller-scale observations from two scintillometers. They found a mismatch of 25%, which can partly be explained by three reasons that also apply here. Firstly, even when EC systems are installed directly next to each other, the observations differ, up to 15% in the case of *ET* (Mauder et al., 2006, 2013). These differences are partly due to large turbulent structures that are not resolved at (sub-)hourly timescales. This makes time-averaged EC observations not by definition representative of the spatial average over heterogeneous surfaces. As these structures may resolve at daily timescales, it may explain the better performance of the bottom-up approach at the daily timescale. Secondly, the footprints are calculated with an analytical model that does not account for surface heterogeneity and 3-dimensional surfaces (more in Section 5.3). Lastly, the patch-scale observations are not necessarily representative of the whole neighborhood scale. In our case, for example, sap flow was measured at six trees that cannot capture the diversity of the trees in the EC footprint. Another example is the low vegetation that experiences shading depending on the location within the canyon.

Still, the physics-driven bottom-up approach yields errors comparable to urban land surface models from a decade ago and only slightly higher than more recent models (Grimmond et al., 2011; Lipson et al., 2023). Most urban land surface models assume the neighborhood flux is the sum of the separate surface covers. Compared to these models, our approach reduces complexity and requires fewer inputs. The found agreement underscores the potential for utilizing surface-specific contributions to decipher *ET* dynamics.

534 In contrast, the top-down approach yields highly unlikely results as the linear sys-  
 535 tem follows mathematics instead of physics. The linear system counteracted high neg-  
 536 ative fluxes with high positive fluxes giving results as extreme as  $-2.0 \times 10^{17}$  and  $1.1 \times 10^6$   
 537  $\text{mm d}^{-1}$ . These effects were omitted from the analysis by excluding negative fluxes, which  
 538 omitted the high fluxes as well due to the linear relations in the equations. Due to these  
 539 direct links, the *ET* contributions contain the errors from the EC observations. How-  
 540 ever, these random errors will cancel out against each other, as we only look at aggre-  
 541 gated results from the top-down results.

542 While open water contributes little to the surface cover, we included this surface  
 543 cover type in our analysis. It cannot be assigned to any of the other surface cover types  
 544 and its inclusion improves the transferability of our methodology. In the top-down ap-  
 545 proach, the low open water fraction caused extremely high or low evaporation fluxes, as  
 546 these could be compensated with relatively small changes in surface cover types with larger  
 547 fractions. In an attempt to eliminate this compensation, we performed the top-down anal-  
 548 ysis with only three surface cover types and three EC systems leaving out the 2-m EC  
 549 at ROTH. As this EC system had the most data gaps, the number of timesteps with re-  
 550 sults rose from 44 to 527. Logically, no contribution from water is inferred. The imper-  
 551 vious contribution rises with 8% (ROTH) and 15% (TUCC) but is still lower than its  
 552 surface fraction. The low vegetation contributes more to *ET* with an increase of 37%  
 553 (ROTH) and 17% (TUCC). Consequently, its *ET* contribution is now higher instead of  
 554 lower than its surface fraction. Only, high vegetation has a lower *ET* contribution drop-  
 555 ping 22% (ROTH) and 24% (TUCC). These changes show mathematical top-down ap-  
 556 proach can give considerably different results by changing the input.

## 557 5.2 Evaporation dynamics per surface cover type

558 Apart from the different *ET* contributions compared to the surface fraction, evap-  
 559 oration evolves differently for each surface cover after rainfall. Impervious surfaces evap-  
 560 orate all water quickly after rainfall, as was also found by (Ramamurthy & Bou-Zeid,  
 561 2014). In contrast, open water sustains evaporation for a longer time. The open water  
 562 evaporation shows a strong daily trend reaching zero during the night. Previous research  
 563 shows the large heat capacity of water dampens the daily trend, which does not go down  
 564 to zero (Jansen et al., 2023). In this study, the daily trend results from the Penman equa-  
 565 tion (Jansen & Teuling, 2020), which was applied given the unavailability of water tem-  
 566 peratures. High and low vegetation show different behavior from each other with the high  
 567 vegetation having a higher initial *ET*. While high vegetation decreases *ET* within days  
 568 after the last precipitation, low vegetation sustains high *ET* rates until soil moisture avail-  
 569 ability is limiting. This soil moisture limitation only occurred towards the end of the sum-  
 570 mer, even though our study year 2019 was relatively dry. The same responses were found  
 571 in other studies (Teuling et al., 2010; van Dijke et al., 2023). High vegetation has a stronger  
 572 stomatal control that enables it to limit transpiration with sufficient available moisture,  
 573 while low vegetation keeps transpiring until it lacks water.

## 574 5.3 Footprint variability and modeling

575 Given these differences in evaporation behavior between surface cover types, the  
 576 surface composition in the footprint influences the EC observations. This changing foot-  
 577 print has to be accounted for to understand *ET* dynamics, as the footprint contribution  
 578 of a particular surface cover may vary as much as 50%. Previously, the relevance of foot-  
 579 prints for *ET* was illustrated by the improved performance when the footprint-weighted  
 580 surface cover was supplied to machine learning models in addition to meteorological ob-  
 581 servations (Vulova et al., 2021, 2023). For other fluxes, such as  $\text{CO}_2$ , footprint model-  
 582 ing has also been shown to help understand flux dynamics (e.g. Velasco et al., 2009; Conte  
 583 & Contini, 2019; Wu et al., 2022).

584 CO<sub>2</sub> sources including directly from humans have been identified and quantified  
 585 by looking at the relation between the CO<sub>2</sub> flux and the surface cover composition equiv-  
 586 alent to Figure 4. For example, Stagakis et al. (2019) find that traffic is an important  
 587 CO<sub>2</sub> source and human respiration accounts for 19% of the CO<sub>2</sub> flux. Human respira-  
 588 tion and perspiration are unlikely to affect our results. In the center of Beijing, the wa-  
 589 ter fluxes from these processes are so small they would account for only 3% of *ET* in Berlin  
 590 (Liu et al., 2022). Given the lower population density of our sites, human respiration and  
 591 perspiration are even lower. Thus, these small water fluxes from these sources are much  
 592 smaller than *ET* and do not influence the results.

593 Footprint modeling is the key that connects the surface to the *ET* in this study.  
 594 The key is however limited by the simplifications of the footprint model. Here, we ap-  
 595 plied the analytical model by Kljun et al. (2015), which generates perfectly symmetri-  
 596 cal footprints. The model does not account for the complexity and heterogeneity of the  
 597 urban morphology. More detailed footprint modeling would provide footprints depend-  
 598 ing on urban morphology, but this would also require more computational resources and  
 599 thus limit the length of the period that can be studied.

#### 600 5.4 Generalizability

601 Here, we studied *ET* in one city during the warm months of a single year, 2019,  
 602 which was a relatively dry year in Berlin. While the climate and year-to-year variabil-  
 603 ity may affect some aspects of the *ET* dynamics, others are likely to be more constant.  
 604 The main aspect we expect to be relatively constant is the evolution of *ET* over a dry-  
 605 down. The impervious surface will evaporate with a short intense peak, open water will  
 606 evaporate more constantly, and vegetation will respond to soil moisture. These general  
 607 patterns may be the same, but the dynamics are altered by site characteristics such as  
 608 plant species, building materials, and water depth. Still, we anticipate this effect to be  
 609 smaller than the differences found between the four surface covers. Apart from site char-  
 610 acteristics, weather conditions control how much each of the surface covers contributes  
 611 to the *ET* (Jansen et al., 2023). The weather conditions determine the water availabil-  
 612 ity (number and length of drydowns), energy availability (radiation and temperature),  
 613 and exchange efficiency (wind and vapor pressure deficit). These conditions will lead to  
 614 changed *ET* dynamics dependent on the season, the climate, and the year-to-year vari-  
 615 ability.

616 The unique 2019 dataset from Berlin allowed us to reconstruct the *ET* signal from  
 617 EC systems. Although relatively common observations are required for the conceptual  
 618 models of the open water and impervious surfaces, the data needed to estimate the evap-  
 619 oration dynamics of both vegetated surfaces is more specialized. These observations in-  
 620 cluded low-level EC observations, tree sap flow, and multiple, continuous soil moisture  
 621 sensors. In most cities, this will not be available. Instead, the vegetated surfaces could  
 622 be modeled with the Penman-Monteith equation (Monteith, 1965). Grimmond and Oke  
 623 (1991) have adapted this equation to urban environments and included the effect of wa-  
 624 ter limitation. As a preliminary analysis, the Penman-Monteith equation was used to  
 625 represent vegetation in the bottom-up approach. This analysis showed that despite an  
 626 overestimation of *ET*, *ET* may be reconstructed with less specialized observations.

## 627 6 Conclusion

628 This study explores the link of neighborhood-scale *ET* to the surface cover at two  
 629 sites in Berlin to estimate the contribution of each surface cover type to *ET*. This link  
 630 is made starting from the *ET* dynamic from the surface cover types reconstructing the  
 631 neighborhood-scale flux (bottom-up) and from four neighborhood-scale fluxes partitioned  
 632 over the surface cover types through a linear system of equations (top-down). We find  
 633 most *ET* originates from vegetation with especially high vegetation evaporating dispro-

634 proportionately more than its surface fraction. Even though impervious surfaces contribute  
 635 less to  $ET$  on long timescales they evaporate substantially after rainfall. Therefore, they  
 636 should not be ignored in urban water management. While both approaches support these  
 637 conclusions, the bottom-up approach proved to be more successful than the top-down  
 638 approach in linking the surface covers at the patch scale to the observations at the neigh-  
 639 borhood scale.

640 We stress the importance of time-dependent EC footprints to understand  $ET$  dy-  
 641 namics. Based on these dynamics, urban land surface models and their evaluation could  
 642 be improved by accounting for the changing footprint. With footprint information, para-  
 643 meters could be dependent on the situation in the current source area. In this way, the  
 644 models would more directly represent what the EC system observes making for a more  
 645 fair and better evaluation.

646 Understanding  $ET$  is crucial in urban water management, for example, to deter-  
 647 mine appropriate vegetation species and irrigation requirements. At the same time,  $ET$   
 648 plays a role in the energy balance and can contribute to the mitigation of heat stress.  
 649 Therefore, the gained insights can support design decisions in city landscapes and ur-  
 650 ban water management to improve the living environment of urban inhabitants.

## 651 7 Open research

652 Spatial datasets are available at the Berlin Digital Environmental Atlas (Senate  
 653 Department for Urban Development, Building and Housing, 2012, 2013; Senate Depart-  
 654 ment for the Environment, Mobility, Consumer and Climate Protection Berlin, 2014).  
 655 Rainfall observations can be accessed at the DWD Climate Data Center (DWD, 2021a).  
 656 Sap flow and soil moisture data are available at the FRED open-access database of IGB  
 657 (Kuhlemann et al., 2020). All other data in this publication is available at 4TU (cita-  
 658 tion at publication).

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 667 in View of Climate Change (UCaHS)” and the research program “Urban Climate Un-  
 668 der Change ([UC]2)”, funded by the German Ministry of Research and Education (FKZ  
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