1 **Optimizing Carbon Cycle Parameters Drastically Improves Terrestrial Biosphere**

2 **Model Underestimates of Dryland Mean Net CO$_2$ Flux and its Inter-Annual**

3 **Variability**

4

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18 **Key Points:**

19 • ORCHIDEE terrestrial biosphere model drastically underestimates dryland mean annual

20   net CO$_2$ fluxes and their inter-annual variability (IAV)

21 • Optimizing phenology, carbon allocation, and respiration parameters are crucial for

22   capturing net CO$_2$ flux mean and IAV

23 • Models need to be optimized against dryland CO$_2$ flux data to achieve accurate

24   predictions of dryland's role in global C cycle variability

Abstract

25

26 Dryland ecosystems occupy ~40% of the land surface and are thought to dominate global carbon

27 (C) cycle inter-annual variability (IAV). Therefore, it is imperative that global terrestrial

28 biosphere models (TBMs), which form the land component of IPCC earth system models, are

29 able to accurately simulate dryland vegetation and biogeochemical processes. However,

30 compared to more mesic ecosystems, TBMs have not been widely tested or optimized against in

31 situ dryland ecosystem CO_2 fluxes. Here, we address this gap using a Bayesian data assimilation

32 system and 89 site-years of daily net CO_2 flux (net ecosystem exchange - NEE) data from 12

33 southwest US Ameriflux sites spanning forest, shrub and grass dryland ecosystems, to optimize

34 the C cycle related parameters of the ORCHIDEE TBM. We find that the default (prior) model

35 drastically underestimates both the mean annual NEE at the high elevation forested mean C sink

36 sites and the NEE IAV across all sites. By testing different assimilation scenarios, we showed

37 that optimizing phenology parameters are particularly useful in improving the model's ability to

38 capture both the magnitude and sign of the NEE IAV. At the high elevation forested sites,

39 optimizing parameters related to C allocation, respiration and biomass and soil C turnover

40 reduces the model underestimate in simulated mean annual NEE. Our study demonstrates that all

41 TBMs need to be calibrated specifically for dryland ecosystems before they are used to

42 determine dryland contributions to global C cycle variability and long-term carbon-climate

43 feedbacks.

1 Introduction

44

45   Terrestrial ecosystems currently take up ~30% of anthropogenic CO_2 emissions, thus

46 acting as a substantial global carbon (C) sink (Fu et al., 2017) and providing a critical reduction

47 in the rate of global warming. However, while we know the magnitude of the global C sink to a

48 good degree of certainty, our knowledge of other components of the global C cycle are more

49 uncertain. One such knowledge gap is which ecosystems, and/or which processes, are driving

50 inter-annual variability (IAV) in land net C uptake (Fu et al., 2017). Improving our

51 understanding of the IAV characteristics of the global terrestrial C cycle is key to being able to

52 forecast the future of the land C sink and long-term biosphere-climate feedback (Cox et al.,

53 2013).

54   Recent studies have pointed to drylands (arid and semi-arid ecosystems) as the dominant

55   driver of global terrestrial C cycle IAV (Ahlström et al., 2015; Poulter et al., 2014). High annual

56   variability in net CO_2 exchange in response to plant-available moisture is observed in site-based

57   flux studies in these regions (Biederman et al., 2017; Cleverly et al., 2016; Haverd et al., 2017;

58   Scott et al., 2015). However, the global terrestrial biosphere models (TBMs) used in these recent

59   C cycle IAV regional attribution studies have often only been extensively evaluated against data

60   in more mesic ecosystems (e.g. (Peng et al., 2015; Piao et al., 2013; Raczka et al., 2013; Schaefer

61   et al., 2012)), although studies have evaluated models against eddy covariance flux data from

62   Australian dryland sites (Haverd et al., 2013a; Whitley et al., 2016). TBM optimization (e.g.

63   parameter calibration) has also only typically been carried out using temperate and boreal site

64   data (Haverd et al., 2013a; Kuppel et al., 2014). Therefore, there remains a relative gap in model

65   benchmarking and optimization using dryland C cycle related data.

66   Model benchmarking and optimization studies that have been performed in dryland

67   regions indicate considerable model-data discrepancies in vegetation dynamics, C and water

68   fluxes (Haverd et al., 2013b; MacBean et al., 2015; Renwick et al., 2019; Trudinger et al., 2016;

69   Whitley et al., 2016; Traore et al., 2014). MacBean et al. (2015) showed that calibrating the

70   phenology parameters of the ORCHIDEE TBM (vAR5) using satellite NDVI at global scales

71   could not account for model errors in semi-arid region seasonal cycle and long-term trends in

72   vegetation dynamics. A recent study by MacBean et al. (in review) has demonstrated that global

73   TBMs participating in the TRENDY v7 model intercomparison project drastically underestimate

74   both the mean annual net ecosystem exchange (NEE) and its IAV at a suite of southwestern

75   (SW) US dryland sites due to weak sensitivity of gross primary productivity (GPP) to changing

76   water availability. This analysis is corroborated by (Renwick et al., 2019) who also showed that a

77   semi-deciduous phenology scheme was necessary to accurately predict the magnitude of GPP in

78   a dryland shrubland. SW US hydrology modeling studies have also suggested that parameter

79   calibration is needed to realistically represent semi-arid water fluxes because the default

80   parameters hamper model performance (Natasha MacBean et al., 2020; Hogue et al., 2005;

81   Unland et al., 1996). Given the lack of model parameter calibration studies that have included

82   dryland sites in their optimizations, it remains to be seen whether model-data discrepancies in

83   dryland ecosystem NEE simulations are due to inaccurate model processes or uncertain

84   parameters. Parameter uncertainty may be higher for dryland ecosystems given parameter values

85 were initially measured in the field and/or optimized for more mesic temperate and boreal
86 ecosystems.

87   To address the gap in dryland site model parameter optimization, and to determine if
88 parameter optimization can account for dryland model-data discrepancies in NEE, we used a
89 Bayesian data assimilation (DA) framework to optimize the photosynthesis, phenology, C
90 allocation and turnover, and respiration parameters of the ORCHIDEE TBM using 89 site-years
91 of daily NEE observations of 12 Ameriflux sites spanning SW US semi-arid grass, shrub and
92 forest ecosystems. Following Biederman et al. (2017) and MacBean et al. (in review), we
93 categorized sites based on their mean annual NEE: US-Vcm, US-Vcp, US-Mpj, US-Fuf, US-Wjs
94 and US-Ses are mostly tree-dominated C sink sites; shrub and grass-dominated sites US-Wkg,
95 US-SRG, US-Seg, US-SRM, and US-Whs "pivot" between a mean annual C sink and source;
96 and the US-Aud grassland is a mean source of C. We used the well-established DA system
97 designed for ORCHIDEE (ORCHIDAS: https://orchidas.lsce.ipsl.fr) (Kuppel et al., 2014;
98 MacBean et al., 2018; Peylin et al., 2016), in which a cost function that represents the misfit
99 between the model and the data – taking into account uncertainty in both – is iteratively
100 minimized using the genetic algorithm (GA; see Methods and Data).

101   Beyond investigating if the DA system could account for model-data discrepancies in
102 dyland NEE simulations, our second objective was to identify which parameters (therefore,
103 which processes) may be responsible for model errors. To address this objective, we performed
104 multiple optimization tests with combinations of parameters related to different model processes
105 in order to identify which processes were most influential in improving the model mean annual
106 NEE and IAV. We focused in particular on which processes are responsible for model failure to
107 capture NEE IAV. Our focus on improving NEE IAV was partly because of the dominant role
108 dryland ecosystems are thought to play in controlling global C cycle IAV, and partly because we
109 expected that, with the exception of sites that are a strong C sink, eddy covariance estimates of
110 mean annual NEE may be impacted by uncertainties in CO_2 flux partitioning. We identified three
111 main groups of parameters: parameters related to 1) phenology; 2) parameters related to
112 photosynthesis; and 3) parameters related to all process calculations that occur after gross C
113 uptake (i.e. C allocation, autotrophic and heterotrophic respiration, biomass and soil C turnover
114 and a scalar on the active soil C pool; hereafter grouped as "post C uptake" parameters). We split
115 the parameters into these three groups because GPP has been shown to be the dominant control

116   on dryland NEE IAV (MacBean et al., in review); therefore, we expected that optimizing

117   parameters related to one or both of the main two processes controlling GPP (i.e., phenology and

118   photosynthesis) will result in the strongest improvements in NEE IAV. However, optimizing all

119   parameters related to processes that occur after gross C uptake can also influence NEE;

120   therefore, we included these parameters as a third category. The parameters included in each

121   assimilation scenario are: P1 - all parameters, including all three phenology, photosynthesis and

122   post C uptake parameter groups; P2 - phenology and photosynthesis parameters; P3 - phenology

123   and post C uptake; P4 - photosynthesis and post C uptake; P5 - phenology parameters only; P6 -

124   photosynthesis only; and P7 - post C uptake only. See **Table 2** for a description of all parameters

125   and to which category they belong.

126   For all assimilation scenarios we compared the prior simulation (before parameter

127   optimization) to the posterior simulations (after parameter optimization, with different parameter

128   groupings for the different assimilation scenarios) by evaluating the simulations against the site

129   data using standard goodness of fit metrics (root mean square error, RMSE and correlation

130   coefficient, r) at daily, monthly and inter-annual timescales. We further attributed what might be

131   causing model-data misfits by decomposing the daily mean squared deviation (MSD) into its

132   component phase, variance and bias contributions. The bias, variance and phase indicate the

133   mean difference in flux magnitude, the mismatch in terms of the magnitude of fluctuations, and

134   the seasonality in flux time series, respectively (Kobayashi & Salam, 2000). All methods and

135   data are described in Section 2 and the results are presented and discussed in Section 3.

136   **2 Methods and Data**

137   2.1 Study sites

138   Twelve semi-arid eddy covariance flux sites in the southwestern US (SW US) have been

139   utilized in this study, with a measurement period ranging between 2003 and 2014. These sites

140   have a range of different vegetation types, climates, elevation and have been described in detail

141   by Biederman et al. (2017), so we only provide a brief description here. We summarize the sites'

142   description, dominant vegetation species, mean climate and corresponding vegetation plant

143   functional types (PFTs), together with the observation period and disturbance history (**Table 1**).

144   The sites are listed consecutively based on their mean annual C balance in Table 1. The major

145   regional IGBP vegetation classes represented include evergreen needleleaf forest, woody

146   savanna, open and closed shrubland, and grassland. These sites typically experience monsoon

147   rainfall during July to October, preceded by a hot, dry period in May and June. The SW US is

148   characterized by water limitation at the annual scale, i.e. potential ET is greater than

149   precipitation. The sites have large spatial gradients in mean annual precipitation (MAP 250–724

150   mm) and temperature (MAT 2.9 to 17.7°C) due to interactions among topography, latitude, wind

151   patterns, and distance from oceans. For further site details, see references in **Table 1** and

152   individual site pages on www.ameriflux.lbl.gov.

153   **Table 1.** Site descriptions, mean climate, observation years and corresponding vegetation plant
154   functional types (PFTs) used in ORCHIDEE optimization. Simulation period corresponds to the
155   period of available site data. PFT acronyms: BS = Bare soil (PFT=1); TeNE = Temperate
156   Needleleaved Evergreen forest (PFT=4); TeBE = Temperate Broadleaved Evergreen forest
157   (PFT=5); TeBD = Temperate Broadleaved Deciduous forest (PFT=6); C4G = C4 grass
158   (PFT=11). Sites are given in order from largest mean annual C sink (US-Vcm) to mean annual C
159   source (US-Aud).

Site ID	Description	Dominant species	IGBP class	PFT fractions	Koppen climate	Elevation (m)	MAP (mm)	MAT (°C)	Period of site data	Disturbance History	Site reference
US-Vcm	Valles Caldera mixed conifer forest	Picea engelmannii, Picea pugens, Abies lasiocarpa var. lasiocarpa, Abies concolor	Evergreen needleleaf forest	100% **TeNE**	Dfb	3042	724	2.9	2007–2012	Harvest 1960s	(Anderson-Teixeira et al., 2011)
US-Vcp	Valles Caldera ponderosa forest	Pinus ponderosa, Quercus gambeli	Evergreen needleleaf forest	100% **TeNE**	Dfb	2501	547	5.7	2007–2014	-	(Anderson-Teixeira et al., 2011)
US-Mpj	Heritage Land Conservancy pinyon-juniper	Pinus edulis, Juniperus monosperma	Savanna	20% BS; 60% **TeNE**; 20% **C4G**	Bsk	2200	423	9.6	2008–2014	-	(Anderson-Teixeira et al., 2011)

US-Fuf	Flagstaff unmanaged ponderosa	Pinus ponderosa	Evergreen needleleaf forest	100% **TeNE**	Csb	2215	607	7.1	2006–2010	Harvest 1910	(Dore et al., 2012)
US-Wjs	Tablelands juniper savanna	Juniperus monosperma, Bouteloua gracilis	Savanna	15% **TeNE**; 85% **C4G**	Bsk	1931	349	10.9	2008–2014	-	(Anderson-Teixeira et al., 2011)
US-Ses	Sevilleta creosote shrubland	Larrea tridentata, G. sarothrae	Open shrubland	20% BS; 55% **TeBE**; 25% **C4G**	Bsk	1610	252	12.6	2007–2014	-	(Petrie et al., 2015)
US-Wkg	Walnut Gulch Kendall grassland	Eragrostis lehmanniana, Bouteloua spp. Calliandra eriophylla	Grassland	60% BS; 3% TeBE; 37% **C4G**	Bsk	1529	386	15.8	2004–2013	Drought 2003-2005, non-native grass replacement 2007 onward, light grazing ongoing	(Scott, 2010)
US-SRG	Santa Rita grassland	Eragrostis lehmanniana	Savanna	45% BS; 11% **TeBD**; 44% **C4G**	Bsh	1292	494	16.7	2009–2014	Mesquite removal 1957, ongoing light grazing	(Scott et al., 2009, 2015)
US-Seg	Sevilleta grassland: burned 2009	Bouteloua eriopoda, Gutierrezia sarothrae, Ceratoides lanata	Grassland	40% BS; 60% **C4G**	Bsk	160	250	12.6	2007–2014	Burned 2009	(Petrie et al., 2015)
US-SRM	Santa Rita mesquite savanna	Prosopis velutina, Eragrostis lehmanniana	Woody savanna	50% BS; 35% **TeBD**; 15% **C4G**	Bsk	1122	421	17.7	2004–2014	Light grazing	(Scott et al., 2009)

| US-Whs | Walnut Gulch Lucky Hills shrubland | Larrea tridentata, Parthenium incanum, Acacia constricta, Rhus microphylla | Open shrubland | 57% BS; 40% **TeBE**; 3% C4G | Bsk | 1376 | 352 | 16.8 | 2008–2014 | Drought 2005-2006 | (Scott, 2010) |
| US-Aud | Audubon grassland | Boutelou agracilis, B. curtipendula, Eragrostis spp. | Grassland | 30% BS; 70% **C4G** | Bsk | 1496 | 348 | 15.7 | 2004–2009 | Burned 2002 | (Krishnan et al., 2012) |

160   2.2 ORCHIDEE terrestrial biosphere model

161   We used the ORCHIDEE (ORganizing Carbon and Hydrology In Dynamic EcosystEms)

162   process-oriented land surface model version 2.2 that has been developed at the IPSL (Institut

163   Pierre Simon Laplace, France). The model is a state-of-the-art mechanistic terrestrial biosphere

164   model (Krinner et al., 2005) and is the land surface component of the IPSLCM5 Earth System

165   Model (Dufresne et al., 2013). The model describes the exchanges of water, carbon, and energy

166   between biosphere and atmosphere at the smallest time scale (30 min), while the slow

167   components of the terrestrial carbon cycle (including carbon allocation, autotrophic respiration,

168   foliar onset and senescence, mortality and soil organic matter decomposition) are computed on a

169   daily to annual basis. Version 2.2 is virtually identical to version 2.0, which is being used in the

170   ongoing Coupled Modeling Intercomparison Project 6 (CMIP6) simulations, but includes few

171   recent bug corrections and code enhancements. It has been updated since the "AR5" version used

172   in CMIP5 (see Krinner et al., 2005) with the following developments: i) an 11-layer mechanistic

173   description of soil hydrology and associated modifications as described in MacBean et al.

174   (2020); ii) addition of a coupled carbon-nitrogen scheme (Vuichard et al., 2019); iii) an

175   analytical solution for the set of equations for photosynthesis, stomatal conductivity and internal

176   CO_2 concentration in the leaf (described in Vuichard et al., 2019), following (Yin and Struik,

177   2009); iv) an update of the soil thermal properties and extension of the soil depth for heat

178   diffusion (Wang et al., 2016); v) a 3-layer snow scheme (Wang et al., 2013); vi) a spatially

179   explicit observation-derived estimate for background albedo and optimized vegetation and snow

180   albedo coefficients;; vii) a new reconstruction of global land cover history and wood harvest

181   accounting following LUH2v2h maps (Hurtt et al., 2020) and PFT maps based the European

182   Space Agency Climate Change Initiative Land Cover product (Poulter et al., 2015).

183   As in most TBMs, the vegetation is grouped into several plant functional types (PFTs),

184   with 14 different types of vegetation plus bare soil in the case of ORCHIDEE v2.2. The original

185   13 PFTs are reported in Krinner et al. (2005). Since ORCHIDEE v2.0 there are now two extra

186   PFTs included: C3 grasses are now split into three groups - tropical, temperate and boreal. The

187   equations governing individual processes are generic with PFT specific parameters, except for

188   the phenology models (see Appendix A in MacBean et al., 2015)). In this study, ORCHIDEE

189   was mainly used in a "grid-point mode" at each site location and forced with the corresponding

190   local 30-minute gap-filled meteorological forcing data. Before performing the optimizations the

191   modelled C stocks were brought to equilibrium in the spin-up phase by cycling the available site

192   meteorological forcing over a long period (1300 years) with the default parameters of the model,

193   which ensures a net carbon flux close to zero over annual-to-decadal time scales.

194   2.3 ORCHIDEE data assimilation system

195   The ORCHIDEE Data Assimilation System (ORCHIDAS) has been described in detail in

196   previous studies (Bastrikov et al., 2018; Kuppel et al., 2014; MacBean et al., 2018; Peylin et al.,

197   2016), and hence we only briefly define the method here. ORCHIDAS uses a variational data

198   assimilation method to optimize the model parameters, accounting for uncertainties regarding the

199   observations, the model, and the prior parameters. It relies on a Bayesian framework with the

200   assumption of Gaussian errors, and the optimized parameters corresponds to the minimization of

201   the following cost function *J(x)* (Tarantola, 2005):

202    $$\boldsymbol{J(x) = \frac{1}{2}[(H(x) - y)^T . R^{-1} . (H(x) - y) + (x - x^b)^T . B^{-1}(x - x^b)]} \qquad (1)$$

203   where *x* represents the parameters and *H(x)* the model contingent on the parameters, and

204   *y* the observations. The cost function contains both the misfit between observations,and

205   corresponding model outputs (first term on the right hand side of Eq. 1), and the misfit between a

206   priori parameter values x_b and optimized parameters *x (second term on the right hand side of the*

207   *Eq. 1 R* is the observation error covariance matrix (including measurement and model errors),

208   and *B* is the prior parameter error covariance matrix. Both matrices (B and R) are diagonal since

209   observation and model errors are assumed to be uncorrelated in space and time, and parameters

210   are assumed to be independent. The cost function is iteratively minimized using the genetic

211   algorithm (GA), which is a meta-heuristic optimization algorithm and follows the principles of

212   genetics and natural selection (Goldberg et al., 1989; Haupt et al., 2004). The GA algorithm has

213   been applied previously with ORCHIDAS tool and described in details by Bastrikov et al.

214   (2018). Briefly, the algorithm works iteratively and considers the vector of parameters as a

215   chromosome and each parameter as a gene on that chromosome. The method fills a set of *n*

216   chromosomes at every iteration, having the starting pool as a randomly perturbed parameter

217   pool. The chromosomes at each subsequent iteration are chosen from randomly selected

218   chromosomes of the previous iteration by either "crossover" or "mutation" process. Santaren et

219   al. (2014) showed that the performance of the algorithm is highly sensitive to its specific

220   configuration and found the best configuration based on computational efficiency after testing

221   different options. Here, we used the same configuration (i.e. number of chromosomes in the pool

222   total number of parameters optimized; the number of iterations is 40; crossover/mutation ratio is

223   4:1; the number of gene blocks exchanged during crossover is 2 and the number of genes

224   perturbed during mutation is 1) applied by Santaren et al. (2014) and Bastrikov et al. (2018). The

225   algorithm does not assume prior knowledge of Gaussian PDFs for the observation and parameter

226   uncertainties. Given we do not fully know the model uncertainty, we set the prior observation

227   uncertainty as the RMSE between the model and the observations following Kuppel et al. (2014).

228   The prior parameter uncertainty is listed in Table S1.

229   The posterior error covariance matrix of the parameters (*A*) can be estimated by:

230   $$A = [H^T R^{-1} H + B^{-1}]^{-1} \qquad\qquad (2)$$

231   This computes error correlations between parameters with the assumption of Gaussian

232   prior errors and linearity of the model in the vicinity of the solution.

233   2.4 Flux measurements

234   At all twelve SW US sites, flux tower instruments collect 30-minutes measurements of

235   meteorological forcing data and eddy covariance measurements of net surface energy and carbon

236   exchanges, which are available from the AmeriFlux data portal (http://ameriflux.lbl.gov).

237   Meteorological forcing data included air temperature and surface pressure, precipitation,

238   incoming long and shortwave radiation, wind speed, and specific humidity. To run the

239   ORCHIDEE model, we partitioned the in-situ precipitation into rain and snowfall using a

240   temperature threshold of 0°C. The site-level meteorological forcing data were gap filled utilizing

241   downscaled and corrected ERA-Interim data following the approach of Vuichard & Papale

242   (2015). Gross primary productivity (GPP) and the ecosystem respiration (R_{eco}) were estimated

243   from the net ecosystem exchange (NEE) via the flux partitioning method described in Biederman

244   et al. (2016). We acknowledge that GPP and R_{eco} are not fully independent data with respect to

245   NEE and are essentially model-derived estimates, but these concerns have been largely discussed

246   in previous studies e.g., Desai et al. (2008). Note that in this study, negative NEE refers to net

247   CO_2 uptake into the ecosystem. In order to exclude the influence of the short-term variations in

248   the fluxes on the model optimization, the daily averaged observations smoothed with a 15-day

249   running mean were used in the assimilation as per Bastrikov et al. (2018).

250   2.5 Parameters optimized

251   The optimized parameters are described in Table S1 with their prior values, prior

252   uncertainty, and upper and lower bounds for different plant functional types based on literature

253   analysis, parameter databases and expert knowledge of the model equations. Prior values are the

254   default parameter values used in all non-optimized ORCHIDEE simulations. In the most past

255   ORCHIDAS studies with previous versions of ORCHIDEE, only subsets of ORCHIDEE C cycle

256   parameters have been optimized (Bastrikov et al., 2018; Kuppel et al., 2012, 2014; MacBean et

257   al., 2015; MacBean et al., 2018; Santaren et al., 2007; Verbeeck et al., 2011). In this study, we

258   considered all possible C cycle related ORCHIDEE parameters to fully explore all sources of

259   parameter uncertainty that is contributing to uncertainties in modeled net and gross CO_2 fluxes.

260   We further allowed weak constraints in the DA system (i.e., large prior parameter bounds, albeit

261   within realistic limits) because the main objective of our study was to determine if parameter

262   calibration can account for model-data errors and to use our assimilation scenario tests to identify

263   which processes are responsible for model-data errors. We selected all 102 parameters and

264   divided them into four classes, controlling the main C cycle and plant physiological processes

265   i.e. photosynthesis, conductance, phenology and post C uptake. This resulted in 31 parameters

266   related to photosynthesis, 42 to phenology, 16 to post C uptake (C allocation, respiration,

267 biomass and soil turnover), and 13 related to conductance. In a preliminary study, we tested at

268 several SW US sites (US-Vcp, US-Mpj, US-Fuf, US-Wkg, US-Whs, US-Seg) the sensitivity of

269 the ecosystem fluxes (NEE, GPP and R_{eco}) when optimizing all model parameters and when we

270 just optimized subsets of the parameters related to each of the main processes. This test showed

271 no significant optimization improvement by adding the conductance related parameters (results

272 not shown here), and thus we did not include those parameters for all final optimizations

273 presented in this study, leaving a total of 89 optimized parameters for each site. Documentation

274 on the parameters can be accessed via ORCHIDEE webpage

275 (https://forge.ipsl.jussieu.fr/orchidee/wiki/Documentation/OrchideeParameters, last access: 04

276 January 2021). The prior uncertainty was set to 40% of the bounds for each parameter following

277 previous ORCHIDAS studies (Kuppel et al., 2012; MacBean et al., 2015).

278 2.6 Assimilation Scenarios

279 We conducted several different assimilation scenarios to identify which processes (and

280 their related parameters) are potentially causing model-data discrepancies (listed in Table 2). We

281 grouped the optimizations based on various parameters set to optimize; therefore, we tested 7

282 assimilation scenarios (P1 – P7): P1 included all 89 parameters, whereas each consecutive

283 scenario (P2 – P7) optimized different subsets of parameters related to each of the main C cycle

284 processes (Table 2). The parameters that were not optimized were set to their default (prior)

285 value. Comparing the P1 to P7 assimilation scenarios allows us to determine which sets of

286 parameters (i.e. specific processes) are contributing most to the improvement in fluxes as a result

287 of the parameter optimizations and therefore provides insight into which model processes may

288 need further modification or development. See Table S1 for groupings of model parameters

289 according to specific processes.

290 **Table 2.** Description of the different assimilation scenarios conducted in this study. The included
291 parameter group(s) and numbers of parameters for each assimilation scenario are given.
292 Parameters of each subgroup are listed in Table S1.

Optimization	Parameters included	Number of parameters
P1	All parameters (Phenology, Photosynthesis and Post C uptake)	~85
P2	Phenology and Photosynthesis	~70

P3	Phenology and Post C uptake	~50
P4	Photosynthesis and Post C uptake	~60
P5	Phenology only	~30
P6	Photosynthesis only	~45
P7	Post C uptake only	~15

293   2.7 Post-optimization analysis

294   We assessed the goodness of fit of the different assimilation scenarios by the mean square

295 deviation (MSD) (in addition root mean squared error, RMSE or correlation coefficients, R and

296 the slope of linear least-square regression). Model evaluation metrics are presented in one of

297 three ways: i) for each site; ii) grouped across all sites; and iii) sites grouped according to their

298 mean net annual CO_2 flux characteristics across the observed time period as in Biederman et al.

299 (2017). For the latter, the net CO_2 "sink" sites are US-Vcm, US-Vcp, US-Mpj, US-Fuf and US-

300 Wjs; the "pivot" sites are US-Ses, US-Wkg, US-SRG, US-SRM, US-Whs, US-Seg; and the

301 "source" site is US-Aud. We followed the approach of Kobayashi & Salam (2000) to quantify

302 the differences between the simulations and observation in terms of bias, variance and phase

303 contribtions to the overall mean squared deviation (MSD). We calculated the MSD between

304 daily model and observed time series and decompose it following the below equation:

305 $$MSD = \frac{1}{n}\sum_{i=1}^{n}(x_i - y_i)^2 = (\overline{x} - \overline{y})^2 + (\sigma_x - \sigma_y)^2 + 2\sigma_x\sigma_y(1 - R) \qquad (3)$$

306   where x is the model and y is the observations, σ is the standard deviation and R is the

307 correlation coefficient.

308   The first term specifies the bias between model simulation and observation (squared).

309 The second "variance" term measures their differences in terms of variability (i.e., the difference

310 between the magnitude of the modeled and observed fluctuations). The third term in Eq. 3

311 generally demonstrates the lack of correlation between model and observations weighted by their

312 standard deviations, which can be deemed a measure of their disagreement in terms in phase

313 (Bacour et al., 2019; Gauch et al., 2003). We further calculated the contribution of each

314 component (bias, variance and phase) to the overall MSD by dividing each component by the

315 total MSD.

3 Results

3.1 Impact of optimization of all parameters (P1) on model net and gross CO_2 fluxes

Across all sites, the prior ORCHIDEE simulations (i.e. before parameter optimization) fail to capture both the mean annual NEE at mean C sink and source sites and the NEE IAV across all sites (**Figure 1a**) - as also seen for all TRENDY TBMs in MacBean et al (in review). Across all sites, optimizing all C cycle-related parameters (phenology, photosynthesis and post C uptake - assimilation scenario P1) with NEE data dramatically increases the ability of the model to capture both the mean C source/sink behavior and the IAV (**Figure 1b**). C sink and source sites show significant improvement in terms of both mean annual NEE and IAV. There is not a strong bias in the model simulations at pivot sites whose mean annual NEE is close to zero; therefore, the optimization results in an improvement mainly in IAV (as represented by the correlation and slope values shown in inset figures in Figures 1 a and b). Improvement of the model-data fit resulting from the assimilation is evident across all sites, with a reduction of daily NEE RMSE between 0.05 to 0.65 $gCm^{-2}d^{-1}$ (**Figure S1**), with a similar reductions in daily GPP and R_{eco} RMSE (Table S2). Moreover, the temporal dynamics are well captured for all the sites: when optimizing all parameters, the median pearson correlation coefficients (R) increase by 0.45, 0.45, and 0.25 for daily, monthly and annual modeled NEE, respectively and posterior median slope values ≥0.5 (**Figure S2a and d)**. GPP temporal dynamics are also much improved by the P1 assimilation with a higher and tighter range in posterior R and slope values than NEE (**Figure S2b and e**). In contrast, there is less improvement in R_{eco} temporal dynamics although the median R and slope values are higher after the optimization (**Figure S2c and f**).
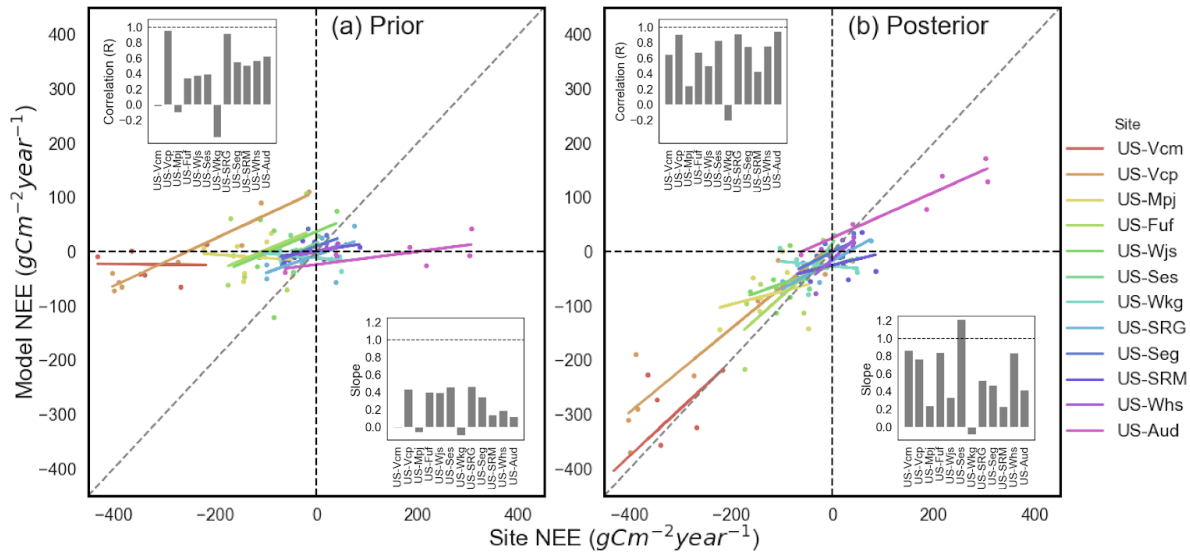
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Figure 1. Comparison between modeled and observed annual NEE when assimilating NEE data and optimizing all phenology, photosynthesis and post C uptake parameters (P1) in the same assimilation. (**a**) Prior annual NEE simulation before parameter optimization, and (**b**) Posterior annual NEE after optimization. The trendline and slope value for the linear regression between the model and observations (bottom right inset figures) is shown for each site, together with their Pearson correlation coefficient, r (top left inset figures). The middle of the trend line should sit on the 1:1 line if the accurate mean annual source/sink behavior for a site is well captured by the model. A slope value close to or equal to 1 demonstrates the model is better at capturing the IAV. Colored points and trend lines represent all twelve sites, ordered from the largest mean sink (US-Vcm) to the largest mean source (US-Aud). The sink sites are: US-Vcm, US-Vcp, US-Mpj, US-Fuf, US-Wjs and US-Ses; the pivot sites are: US-Wkg, US-SRG, US-Seg, US-SRM and US-Whs; and the only source site is: US-Aud.

Across the majority of the sites, the prior model simulates a depressed seasonal NEE amplitude and/or is unable to capture the observed bi-modal seasonality (**Figure 2**). The NEE amplitude and bi-modal seasonality generally improved when optimizing all parameters (blue curves in **Figure 2**), although the posterior simulations struggle to reach the exact magnitude of the spring and monsoon NEE troughs (net CO_2 uptake) for several sites (e.g. US-Mpj, US-Wjs, US-Ses, US-Seg, US-Wkg and US-Whs). Accurately capturing the seasonal peaks and troughs is important for replicating observed NEE IAV because variability in summer monsoon season fluxes are the dominant driver of NEE IAV (MacBean et al., in review). While posterior seasonal NEE peaks and troughs are generally well captured, the assimilation of NEE alone often fails to capture the correct peaks in gross CO_2 fluxes (**Figure S3**), likely due to compensating errors in

361   both GPP and R$_{eco}$. At the C source site (US-Aud) the model also fails to simulate the accurate

362   peaks in springtime net carbon release (**Figure 2**). This is due to the fact that at US-Aud, TBMs

363   tend to overestimate spring GPP and underestimate the earlier rise in spring R$_{eco}$ (**Figure S3**).

364   The optimization only partially corrects these model biases, suggesting that other missing

365   processes may ultimately be responsible for the model-data misfit (such as disturbance following

366   a fire that occurred at the site in 2002, which is not implemented in the current version of

367   ORCHIDEE).

368

369   **Figure 2.** Mean monthly NEE seasonal cycles for each site comparing prior (red curve) and
370   posterior (blue curve) ORCHIDEE simulations with observations (black curve). Posterior
371   simulation after assimilation of NEE data and optimization of all parameters: phenology,
372   photosynthesis and post C uptake (P1). The sites are listed in order from largest mean annual C
373   sink (US-Vcm) to mean annual C source (US-Aud).

374

375   Decomposing the daily NEE MSD between model and observations into bias, variance

376   and phase components shows that across all sites, all three components contribute to prior NEE

377   model-data discrepancies (**Figure 3a** left of vertical dashed line). The prior daily NEE MSD at

378   the C sink sites are dominated by both phase and bias components (**Figure 3a** top panel). The

379   fact that sink site MSD is also dominated by bias is unsurprising given that at those sites the prior

380   model does not capture the mean annual C sink (**Figure 1a**). Note that, if we decompose the

381   *annual* NEE MSD into the constituent bias, phase and variance components then bias

382   overwhelmingly dominates the MSD at sink (and source) sites given their large underestimate of

383   mean annual NEE (**Figure S4** top and bottom rows). In contrast, at the C pivot and source sites,

384   the highest contribution to the prior daily NEE MSD is from the phase component (**Figure 3a**

385   middle and bottom panel), indicating that the default model does a poor job of representing the

386   timing of dryland C cycle related processes. Across all sites, optimizing all parameters (P1)

387   dramatically reduces the bias, variance and phase components of the daily NEE MSD, with

388   phase remaining the strongest contributor to daily NEE MSD (**Figure 3a** right of dashed line).
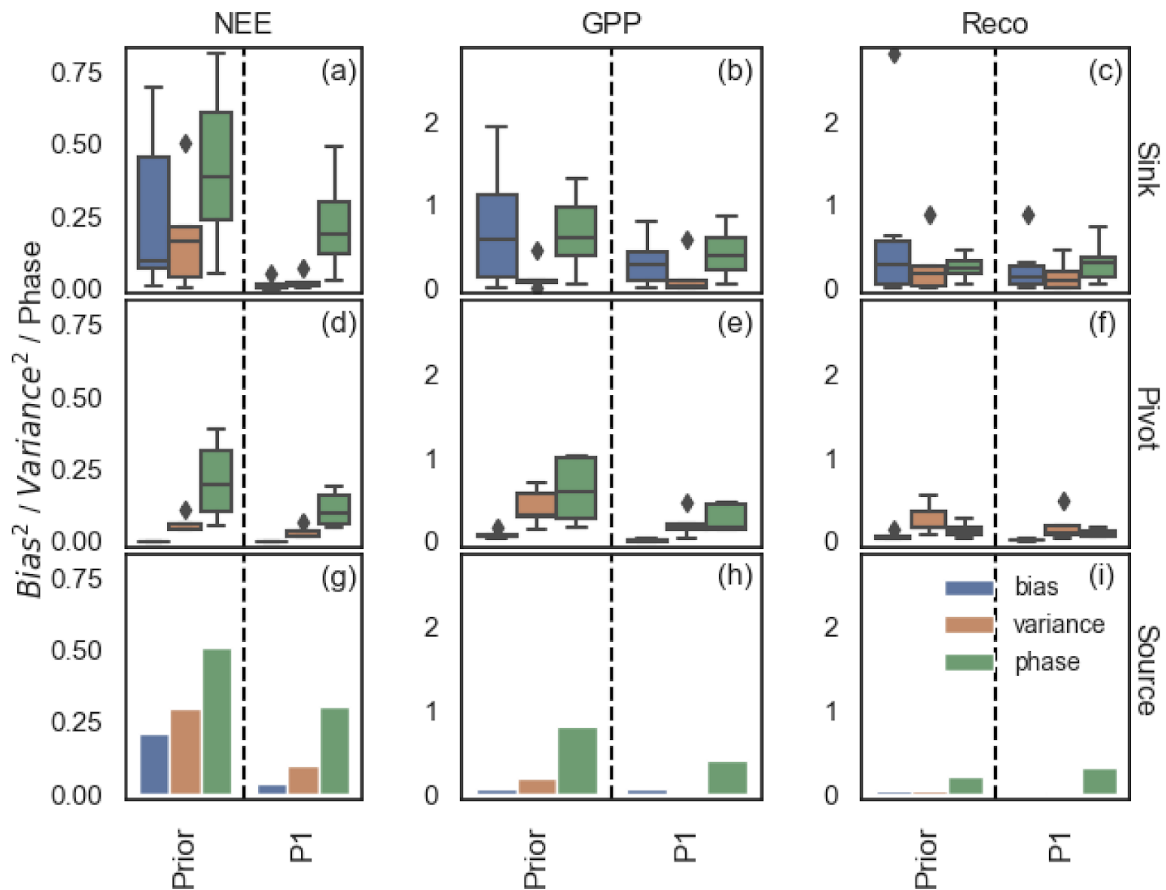

389

Figure 3. Daily NEE, GPP and R$_{eco}$ mean square deviation (MSD) decomposition into bias, variance, and phase between simulations and observations for assimilating NEE observations and optimizing all phenology, photosynthesis and post C uptake parameters (P1). Blue, orange and green boxplots for bias, variance and phase components, respectively. Different rows separate the sites as sink (a-c), pivot (d-f) and source (g-i) based on total annual C flux. The sink sites are: US-Vcm, US-Vcp, US-Mpj, US-Fuf, US-Wjs and US-Ses; the pivot sites are: US-Wkg, US-SRG, US-Seg, US-SRM and US-Whs; and the source site ia: US-Aud. The x axes display the optimization scenarios (Prior and P1). The box whiskers show the spread of bias, variance and phase for all 12 sites considered in this study. The bias, variance and phase indicate the mean difference in flux magnitude, the mismatch in terms of flux fluctuation magnitude scales with the mean seasonal amplitude, and the seasonality in flux time series, respectively. Note that the y axis limits for both gross fluxes (GPP and R$_{eco}$) are the same.

As for the NEE, bias and phase are the dominant contributors to prior daily GPP MSD for the sink sites (left of vertical dashed line in **Figure 3b**), and phase only for the pivot and source sites (**Figures 3e and h**) For R$_{eco}$, a different MSD component is dominant depending on the mean C behavior of a site: bias dominates the prior daily R$_{eco}$ MSD at the sink sites, variance at the pivot sites, and phase at the source sites (**Figures 3c, f and i**). Overall, assimilating NEE data in the P1 assimilation scenario reduces all gross CO$_2$ flux MSD components (right of dashed line in **Figure 3** middle and left columns), with phase remaining the strongest contributor to daily gross CO$_2$ flux MSD. However, unlike for the NEE, at the C sink sites phase *and* bias remain strong contributors to posterior GPP MSD (**Figure 3b**).

3.2 Impact of different processes (assimilation scenarios) on optimization results

Across all sites, modeled annual and seasonal NEE are improved the most in the P1 assimilation scenario compared to the other assimilation scenarios (P2 to P7), although all scenarios result in some improvement (**Figures S5, S6a and d,** and seasonal cycles in **Figure S7**). In general, there is less improvement in R$_{eco}$ compared to NEE and GPP (**Figure S6**). Comparing the MSD decomposition results for the various assimilation scenarios (P1-P7) can help to identify which processes may be causing the prior model-discrepancies in mean annual NEE and NEE IAV. At the source and sink sites, the bias component (blue bars in **Figure 4a and c**) is reduced dramatically by all optimization tests that include the post C uptake parameters related to C allocation, respiration, and aboveground biomass and soil C turnover (P1, P3, P4 and P7). For the sink sites, assimilation scenarios that also include photosynthesis parameters (P2 and

424   P6) also result in a strong reduction in bias. This decrease in mean bias is also shown by the fact

425   that the midpoints of the linear regression trendline between model and observations at forested

426   sink sites (US-Vcm, US-Vcp, US-Mpj, and US-Fuf) and low-elevation source site (US-Aud)

427   with optimization scenarios P1 to P4, P6 and P7 parameters all lie much closer to the 1:1 (grey
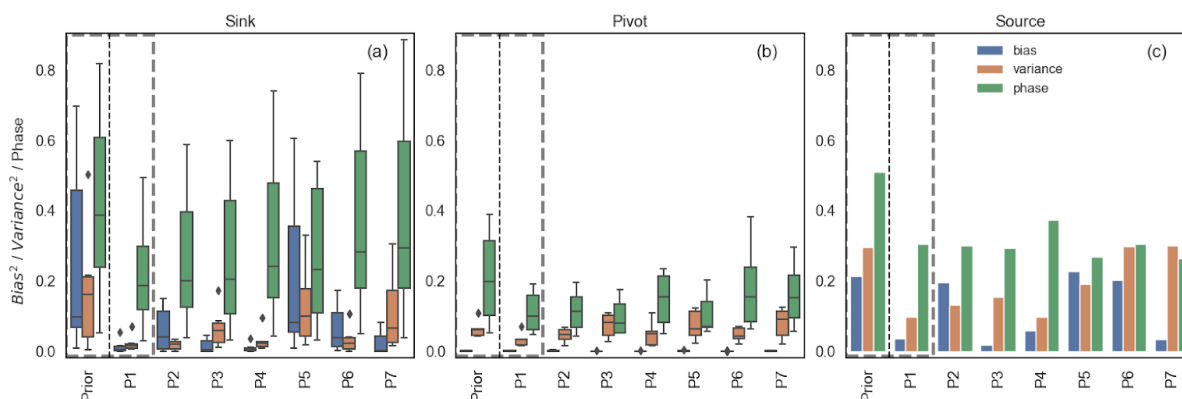
428   dashed) line compared to P5 (**Figure S5**).

429

430   **Figure 4.** Daily NEE MSD decomposition into bias, variance, and phase components when
431   assimilating NEE observations for different assimilation scenarios (P1-P7). Different panels
432   separate the sites as sink (a), pivot (b) and source (c) based on total annual C flux. The C sink
433   sites are: US-Vcm, US-Vcp, US-Mpj, US-Fuf, US-Wjs and US-Ses; the C pivot sites are: US-
434   Wkg, US-SRG, US-Seg, US-SRM and US-Whs; and the C source site is: US-Aud. The grey
435   dashed boxes highlight results repeated from **Figure 3(a,d,g)** to have better comparison of
436   different process parameters side-by-side. The parameters included in each optimization are: P1:
437   all parameters; P2: phenology and photosynthesis; P3: phenology and post C uptake; P4:
438   photosynthesis and post C uptake; P5: phenology; P6: photosynthesis and P7: post C uptake. The
439   boxplots show the median and interquartile range of the bias, variance and phase across all 12
440   sites considered in this study. US-Aud is the only C source site; therefore, the barplots in (c)
441   show the bias, phase, and variance components of the MSD for that one site. The bias, variance
442   and phase indicate the mean difference in flux magnitude, the difference in the magnitude of flux
443   variations, and the difference in the correlations weighted by the standard deviations,
444   respectively (see Methods).

445

446   Across all sites the difference in phase between the model and observations (green bars in

447   **Figure 4**), which, as already noted, is the largest contribution to the prior NEE MSD across all

448   sites, is mostly reduced by assimilation scenarios that include phenology parameters (i.e. P1, P2,

449   P3 and P5). The P4 assimilation (photosynthesis and post C uptake parameters) also does well in

450   reducing phase contributions to NEE MSD at forested C sink sites (Figure 4a). However, the

451   phase component is not reduced as much as the bias in any of the assimilation scenarios; thus, for

452 all sites and all assimilation scenarios the phase remains the largest component of the posterior

453 daily NEE MSD (**Figure 4**). Including parameters related to photosynthesis or post C uptake

454 with the phenology parameters (i.e. assimilation scenarios P2 and P3) helps to slightly reduce the

455 phase discrepancy at sink sites compared with phenology parameters alone (P5) (as seen above

456 for the improvement in slope values at the sink sites). Examining the spread in slope and R

457 values across all sites, we see that the annual variability (median slope and R values) is improved

458 the most for assimilation scenarios with at least two parameter sets (P1 to P4 - **Figure S6a and**

459 **d**). The persistence of phase as the dominant component of the posterior daily NEE suggests

460 further model improvement in processes related to dryland vegetation temporal dynamics (e.g.

461 phenology and all associated processes) is needed before TBMs can correctly reproduce NEE

462 seasonality and IAV.

463 The variance component of the daily NEE MSD (orange bars in **Figure 4**), which also

464 shows a modest contribution to daily NEE MSD at the sink and source sites, is mostly reduced at

465 the sink sites with assimilation scenarios that include photosynthesis parameters (i.e. P1, P2, P4

466 and P6). At US-Aud the variance component was reduced most by assimilation scenarios that

467 included two or more sets of parameters (i.e. P1 - P4) (**Figure 4c**).

468 While the post C uptake parameters are key for reducing bias in forested sink site NEE,

469 biases in GPP and R_{eco} at these sites are reduced by optimizing photosynthesis parameters (P1,

470 P2, P4, and P6 - blue boxes **Figure S8b and c**). The GPP and R_{eco} bias components at the sink

471 sites are not reduced as strongly as NEE biases for any assimilation scenario; thus, bias remains a

472 key contributor to posterior gross CO_2 flux MSD. Similarly to NEE, parameter subsets that

473 include phenology parameters (P1, P2, P3 and P5) are key for reducing the daily GPP MSD

474 phase component at pivot sites (green boxes in **Figure S8e**; however, in contrast with the NEE

475 results, at sink sites the GPP phase component tends to be reduced by all assimilation scenarios

476 except P7 (see also median GPP slope and R values in **Figures S6b and e**). With the exception

477 of P1 and P2 for GPP, the GPP and R_{eco} variance components are not reduced much by any of

478 the assimilation scenarios and remain a considerable component of the MSD for both GPP and

479 R_{eco} at the pivot sites, and for R_{eco} at the sink sites (**Figures S8b,c,e,f**). We note that the GPP and

480 R_{eco} reductions in MSD components tend to be similar, suggesting model-deficiencies in R_{eco} are

481 mainly influenced by those in GPP. Addressing GPP model-data deficiencies is therefore a high

482 priority.

483   3.3 Constraint on parameters

484   For all assimilation scenarios, we found significant parameter deviations from prior

485   values for numerous phenology, photosynthesis and post C uptake related parameters (**Figure 5**),

486   which is consistent with the fact that all parameter subsets are needed to improve model mean

487   annual NEE and IAV. Parameter deviation was calculated using the difference between the

488   posterior and prior parameter value normalized by the total parameter variation used in the

489   optimization. Finally, the median value was taken as the mean deviation from all 12 sites. We

490   did not find that parameters deviate more, or the uncertainty reduction (calculated as 1 –

491   (posterior parameter uncertainty / prior parameter uncertainty)) was much different, when only

492   one subset or two parameter subsets were included in the optimization instead of all three (e.g.

493   cf. P2 with P1), although posterior values are different for each assimilation scenario (**Figure 5**).

494   In particular, most of the post C uptake parameters deviate strongly from the prior median

495   deviations (>20% of total parameter bound). There are also significant uncertainty reductions

496   (>50%) for most of the parameters which show strong deviations from their prior value: 10 for

497   phenology (out of 42), 7 for photosynthesis (out of 31) and 7 for post C uptake (out of 16)

498   (**Figure 5**). The error correlations between the estimated parameters are usually minimal except

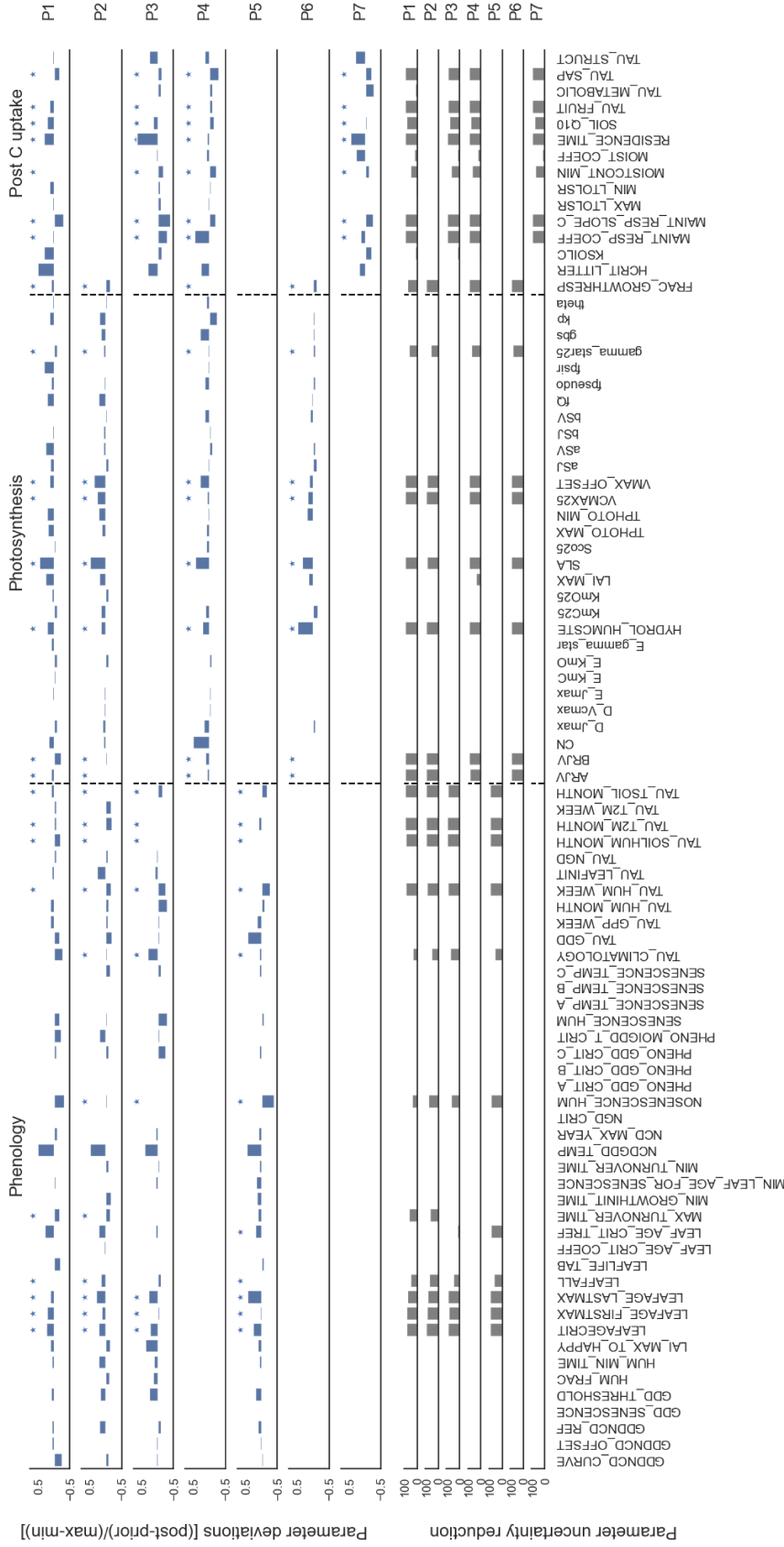499   between post C uptake parameters (see example for one site in **Figure S9**).

500

502   Certain phenology parameters are important across all assimilation scenarios: i)

503   parameters related to leaf age; ii) a parameter related to the critical temperature threshold for the

504   start of deciduous shrub leaf growth (ncdgdd_temp); iii) moisture thresholds that govern C4

505   grass senescence (nosenescence_hum); and iv) various parameters that control the time scales

506   used in phenology schemes (e.g. tau_climatology, tau_hum_week) (**Figure 5**). The phenology

507   models are highly dependent on such empirical parameters, which likely need to be optimized

508   for each site. Key photosynthesis related parameters are SLA, parameters involved in the

509   calculation of $V_{c,max}$ (the maximum carboxylation rate, which has been shown to be a highly

510   sensitive model parameter in previous studies, e.g. Kuppel et al., 2014), and the parameter that

511   represents the root profile in the empirical calculation of leaf water stress (hydrol_humcste),

512   which downregulates photosynthesis and stomatal conductance in the dry season. The most

513   important post C uptake parameters are fairly similar across assimilation scenario tests, and are

514   related to: i) the calculation of the maintenance respiration as a fraction of biomass; ii)

515   aboveground biomass residence time and various turnover rates for biomass and litter pools; and

516   finally, iii) the Q10 parameter involved in the temperature dependence of soil C decomposition

517   (**Figure 5**).

518   **4 Discussion and Conclusions**

519   In this study, we have shown that it is possible to account for model discrepancies in both

520   the mean annual NEE and NEE IAV at a range of semi-arid SW US sites via optimization of C

521   cycle parameters within a Bayesian DA framework. We used weak prior constraints (i.e. large

522   prior parameter bounds) to give the assimilation the maximum chance to correct any model

523   errors. Our goal was not to identify the ideal "correct" set of C cycle parameters for capturing

524   semi-arid vegetation and C cycle dynamics, but rather to identify whether, within the current

525   model representation, we could account for model-data mismatches. Looking at the individual

526   parameter plots for the P1 assimilation scenario (**Figure S10)**, we find that at some sites several

527   posterior parameters are "edge-hitting" (e.g. soil Q10). Given we chose weak prior constraints in

528   the assimilation, the fact that some posterior parameters are hitting their bounds suggests that the

529   optimization may be aliasing model structural error onto the parameters (as demonstrated in

530   MacBean et al., 2016) and/or that the model cannot improve further via parameter optimization.

531   This suggests that further model developments are likely needed to address structural

532   uncertainties and missing processes.

533   Hypotheses as to which processes might be responsible for model inability to capture

534   semi-arid CO_2 fluxes are numerous and will take time to explore fully (MacBean et al., in

535   review). Here, we aimed to speed up that process by using the different assimilation scenarios as

536   tests of which parameter sets (and therefore, which processes) may be responsible for model

537   errors. The assimilation with all C cycle and vegetation parameters (P1) performed the best in

538   terms of correcting underestimates in modeled mean annual NEE and IAV. However, the

539   additional assimilation scenarios (P2 to P7) further demonstrated that phenology parameters are

540   likely key for improving semi-arid ecosystem NEE IAV. Issues with semi-arid phenology in

541   TBMs have been documented elsewhere (Traore et al., 2014; Dahlin et al., 2015; MacBean et al.,

542   2015; Renwick et al., 2019; Whitley et al., 2016; Teckentrup et al., in review). Further evidence

543   for inadequate TBM phenology schemes comes from MacBean et al. (2020), who noted that

544   while the ORCHIDEE model can capture evapotranspiration (ET) fluxes extremely well, even

545   without parameter optimization, the model simulates a delayed increase in transpiration/ET

546   (T/ET) ratios during the summer monsoon when compared to two independent T/ET estimates.

547   This suggests that the model is getting ET right for the wrong reasons – i.e. the partitioning of

548   ET into its component fluxes of T and bare soil evaporation is incorrect. This lagged response of

549   T to increasing rainfall is consistent with the results of MacBean et al. (in review) who found

550   across a suite of TBMs (TRENDY v7) too weak ecosystem-scale water use efficiency (WUE) –

551   i.e. a too weak response of GPP to increasing ET – during the monsoon was likely the cause of

552   their inability to capture NEE IAV. Put simply, the models simulate too weak a response of

553   vegetation growth to pulses of moisture availability. Thus, the evidence from all these studies,

554   including our results presented here, is pointing to issues with phenology, plant hydraulics,

555   and/or the fraction of vegetation prescribed in the model. As noted by MacBean et al. (2020), the

556   static PFT fractions prescribed in the models likely prevent monsoon season growth of summer

557   annual C4 grasses in the interstitial bare soil patches. Errors in PFT fractions in sparsely

558   vegetated regions have also been shown to propagate into large model errors in simulated

559   carbon, water and energy fluxes (Hartley et al., 2017). The optimization of numerous phenology

560   parameters with weak constraints in this study could be partially accounting for such a model

561   error. But it is also possible that this issue of static PFT fractions explains even the posterior

562 model's inability to capture peak GPP fluxes for some sites (see Section 3.1), and the fact that in
563 the posterior simulations, the phase remains the strongest contribution to the NEE MSD.

564   The same Bayesian DA system was used by MacBean et al (2015) to correct phenology
565 model issues in a previous version of ORCHIDEE that was nonetheless identical in its
566 representation of phenology. However, while they were able to correct the seasonal leaf
567 dynamics in temperate and boreal ecosystems, they found the parameter optimization was unable
568 to correct for phenology model issues in semi-arid ecosystems. While the data they used were
569 different – normalized difference vegetation index (NDVI) from the MODIS satellite instrument
570 as opposed to the flux tower NEE used here – they also used stronger prior constraints and fewer
571 phenology parameters, suggesting that the additional degrees of freedom in the assimilations in
572 this study (from weaker prior constraints and a greater number of phenology parameters) may
573 have resulted in the improvements from the parameter optimization. Still, as we noted above, the
574 combination of weak prior constrains and edge-hitting posterior parameters suggests the
575 assimilations are accounting for other structural errors in the model, and phase errors remain a
576 strong source of NEE MSD even after optimization. As also noted, the phenology schemes in
577 these models are highly dependent on a number of empirical parameters that require site
578 calibration and which were typically not developed for dryland ecosystems. Future developments
579 in this area should take account of the variety of different strategies in dryland plants for dealing
580 with water stress (Smith et al., 2012).

581   MacBean et al. (in review) also presented a range of other hypotheses as to which
582 processes might be causing model errors in capturing semi-arid phenology other than the need to
583 represent summer annual C4 grass fractional cover and phenology, including: the lack of
584 drought-deciduous shrub phenology schemes in TBMs; the lack of deep tap shrub and tree roots
585 that draw up groundwater needed for growth during drier periods; and the lack of dynamic root
586 growth as moisture becomes more available. Future studies need to systematically test all of the
587 proposed hypotheses to determine which, if any, can explain the observed model-data
588 discrepancies.

589   Our assimilation tests also showed that so-called "post C uptake" parameters related to
590 maintenance respiration, biomass and litter turnover, and soil C decomposition are mainly
591 responsible for reducing the strong model underestimate of mean annual NEE, particularly at the

592 higher elevation forested C sink sites. We did not focus on correcting mean annual NEE, and

593 instead focused more on correcting errors in NEE IAV, because the variability in eddy

594 covariance measurements of NEE are more trusted than the absolute values due to errors in flux

595 partitioning. Furthermore, for the semi-arid sites that pivot between a C source and sink, their

596 mean sink versus source behavior may be a function of a time period involved. In particular, the

597 only mean C source site (US-Aud) is likely a source because of a fire in 2002 from which the site

598 was still recovering during the measurement period. As discussed, we know that even TBMs that

599 include wildfire modules will likely not reproduce the specific impacts of an individual fire.

600 Nevertheless, while we do not focus on the C source site, we do know that the high elevation

601 forested sites in this study are consistently sinks of C, even during the drought period that has

602 been affecting the SW US for most of this century (Scott et al., 2015). It is important that we are

603 able to capture this dryland forested site C sink, particularly given these ecosystems have been

604 shown to contribute to long-term trends in the global C cycle (Ahlstrom et al., 2015). Drylands

605 are vulnerable to future increases in drought, which may reduce the C sink (Bodner and Robles,

606 2017). On the other hand, drought impacts on dryland vegetation could be mitigated by increases

607 in WUE and vegetation growth under elevated CO_2 (e.g. Donohue et al., 2013). Thus, it is an

608 important contribution that parameter optimizations presented here can account for these biases

609 in modeling C sinks at high elevation forested sink sites. MacBean et al. (in review) postulated

610 that TRENDY TBM underestimates in mean annual NEE at these sites was due to

611 underestimates in spring GPP, possibly due to issues with model snow melt not providing

612 enough moisture for spring growth. In constrast, the results presented here suggest that the issues

613 at the high elevation forested sink sites may be more linked to processes that occur after the

614 gross uptake of CO_2, such as maintenance respiration, biomass turnover, and temperature

615 limitation on soil C decomposition (**Figure S11**). If true, it may be that the processes in TBMs

616 can accurately capture dryland forested site photosynthesis if the parameters are better adapted

617 for dryland PFTs, which simply requires more careful calibration across a range of dryland forest

618 sites.

619   As discussed in the introduction, we have focused on correcting parameters related to

620 GPP because MacBean et al. (in review) found that GPP, and particularly summer monsoon

621 season GPP, is the dominant driver of NEE IAV. We also are obliged to focus on GPP

622 parameters because the number of model parameters is higher for GPP. In a follow up study, we

623   are assessing how the number of parameters linked to each different process affects the ability of
624   the optimization to correct for errors in those processes. We may find, for example, that the sheer
625   number of parameters related to phenology that are included here results in those parameters
626   being the more important for correcting NEE IAV. This then becomes an issue of wider model
627   development because we can only include in the optimization that are in the model. Still, the fact
628   that the relatively few "post C uptake" parameters included in the assimilation tests carried out in
629   this study can account for biases in mean annual NEE suggests that the number of parameters
630   linked to each process does not prevent us from identifying which set of parameters (and
631   processes) are mostly causing model-data discrepancies. It is still possible that those parameters
632   are accounting for other structural errors in the model, as we have discussed above. The specific
633   DA configuration (e.g. type of data included – e.g. net or gross CO_2 fluxes, the number of
634   parameters included, and to which processes they are related) can lead to different posterior
635   values and degree of improvement in model-data fit. Therefore, further tests of different DA
636   configurations and optimizations at other locations are needed to explore the potential of
637   Bayesian DA systems for quantifying and reducing error in semi-arid ecosystem C fluxes. While
638   we have particularly highlighted one key area of the model that needs improvement (dryland
639   phenology schemes and associated processes related to plant water availability), we have also
640   shown that all C cycle model parameters and processes in semi-arid ecosystems need either
641   optimizing or further development by TBM groups. Only by addressing these issues will we be
642   able to reliably use these models to accurately simulate dryland contributions to IAV and long-
643   term trends in the global C cycle.

644

660   **Code availability**

661   The ORCHIDEE model is under a free software license (CeCILL; see

662   http://www.cecill.info/index.en.html) and the source code is visible here:

663   https://forge.ipsl.jussieu.fr/orchidee/browser/tags/ORCHIDEE. The ORCHIDEE model code is

664   written in Fortran 90 and is maintained and developed under an SVN version control system at

665   the Institute Pierre Simon Laplace (IPSL) in France. The ORCHIDAS code is currently in the

666   process of being put on a GitHub repository but for now it is available on request to

667   vladislav.bastrikov@lsce.ipsl.fr.

668

669   **Data availability**

670   Meteorological forcing data and eddy covariance measurements of net surface energy and

671   carbon exchanges at 30-minutes intervals are available from the AmeriFlux data portal

672   (http://ameriflux.lbl.gov). The model outputs from ORCHIDEE simulations and post-processing

673   python scripts for manuscript figures and tables are freely available in a Git repository

674   (https://github.com/kashifmahmud/SW_US_semiarid).

675

676   **References**

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