

1                                   **Accelerated ocean acidification in the Tsugaru Strait**  
2                                   **by an intensified Tsugaru Warm Current**

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21                  **Key Points**

- 22                  • Acidification in the eastern part of the Tsugaru Strait is found to have advanced  
23                  considerably throughout its depth during 2012–2019
- 24                  • Acidification is occurring at faster rates than reported elsewhere and faster than expected  
25                  for oceanic uptake of anthropogenic CO<sub>2</sub>
- 26                  • The accelerated acidification is caused by enhanced increase of dissolved inorganic  
27                  carbon, owing to a strengthened Tsugaru Warm Current

28  
29

30 **Abstract**

31 The oceanic uptake of anthropogenic CO<sub>2</sub> has resulted in acidification in surface and subsurface  
32 waters throughout the global ocean. We initiated acidification monitoring in the eastern part of  
33 the Tsugaru Strait, which connects the Sea of Japan and the North Pacific. Annual mean pH and  
34 CaCO<sub>3</sub> saturation state decreased considerably throughout all depths at rates of 0.0030–0.0051  
35 yr<sup>-1</sup> and 0.017–0.036 yr<sup>-1</sup> during 2012–2019, respectively. These rates of decrease are faster than  
36 those caused by increasing atmospheric CO<sub>2</sub>, and faster than those observed during previous  
37 research in the Sea of Japan and the North Pacific. The accelerated acidification is attributed to  
38 enhanced increase in dissolved inorganic carbon caused by elevated mixing of the upper and  
39 deeper waters from the Sea of Japan at the western part of the strait by the strengthening of the  
40 Tsugaru Warm Current.

41

42 **Plain Language Summary**

43 Approximately 30% of the total amount of CO<sub>2</sub> released to the atmosphere by human activities  
44 has accumulated in the global ocean. This oceanic uptake of CO<sub>2</sub> has resulted in ocean  
45 acidification. In coastal waters the acidification affects marine calcifying organisms, thus coastal  
46 ecosystems may be more vulnerable to acidification than the open ocean. To examine the extent  
47 to which acidification has advanced in the eastern part of the Tsugaru Strait, which connects the  
48 Sea of Japan and the North Pacific, we initiated a time-series observation of acidification.  
49 Acidification is found to have advanced considerably across the whole depth during 2012-2019  
50 at a rate faster than that caused by increasing atmospheric CO<sub>2</sub>, and at the highest rates observed  
51 during previous research in the North Pacific Ocean and the Sea of Japan. The rapid acidification  
52 is found to be attributable to the enhanced rate of increase of dissolved inorganic carbon caused

53 by elevated mixing of the upper and deeper waters from the Sea of Japan at the western part of  
54 the strait by the strengthening of the Tsugaru Warm Current. In other straits that are connected to  
55 the open ocean, the strengthening of their throughflow may also accelerate acidification.

56

## 57 **1 Introduction**

58 Ocean absorption of the CO<sub>2</sub> released to the atmosphere by human activities (i.e.,  
59 anthropogenic CO<sub>2</sub>) results in acidification of surface waters as the result of the increase in  
60 hydrogen ion (H<sup>+</sup>) concentration. The oceanic uptake of anthropogenic CO<sub>2</sub> has decreased ocean  
61 pH by 0.1 since the beginning of the industrial era (Intergovernmental Panel on Climate Change,  
62 2013). Ocean acidification also lowers the CaCO<sub>3</sub> saturation state ( $\Omega$ ) by decreasing carbonate  
63 ion (CO<sub>3</sub><sup>2-</sup>) concentration with concomitant formation of bicarbonate ions. The carbonate  
64 chemistry change affects marine calcifying organisms with carbonate shells and skeletons,  
65 biogeochemical cycling of nutrients, and ecosystems (e.g., Doney et al., 2009; Sugie et al.,  
66 2018).

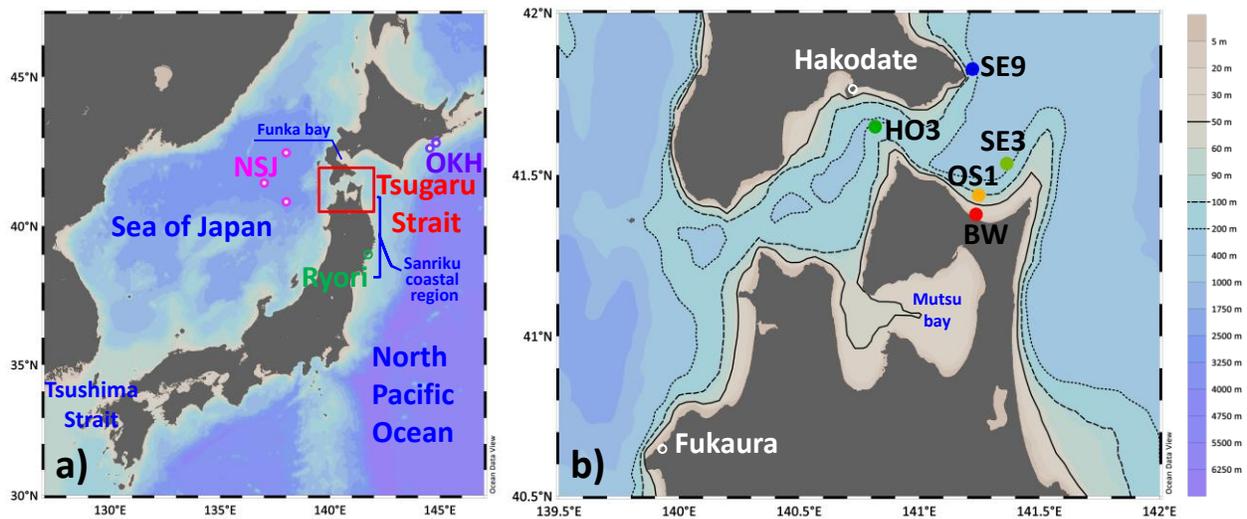
67 In the open ocean, the acidification of surface water has been well documented with time-  
68 series data at CO<sub>2</sub> monitoring sites (e.g., Astor et al., 2013; Bates et al., 2012; Currie et al., 2011;  
69 Dore et al., 2009; González-Dávila et al., 2010; Olafsson et al., 2010; Ono et al., 2019; Wakita et  
70 al., 2017). The rates Olafsson of decrease in surface water pH are mostly within the range of  
71 0.0013–0.0024 yr<sup>-1</sup>, as expected from the response to the release of anthropogenic CO<sub>2</sub>.  
72 Furthermore, in coastal waters, acidification affects marine calcifying organisms; for example,  
73 causing mass death of oyster larvae (e.g., Feely et al., 2008) and damage to the carapaces of  
74 larval Dungeness crabs (Bednaršek et al., 2020) along the continental shelf from central Canada  
75 to northern Mexico. The serious threats that coastal calcifying organisms have already faced  
76 suggest that coastal ecosystems may be more vulnerable to ocean acidification than the open  
77 ocean.

78 Japanese coastal waters are also acidified by the oceanic uptake of increased  
79 anthropogenic CO<sub>2</sub> in the atmosphere (Chen et al., 2017; Ishii et al., 2011; Ishizu et al., 2019;

80 Lui & Chen, 2015). In this study, we focus on the acidification of coastal water at the Tsugaru  
81 Strait. The strait is 100 km in length, 20–40 km in width, and has shallow sills (~130 m depth)  
82 near its western part (Figure 1). The Tsugaru Warm Current (TWC) flows eastward from the Sea  
83 of Japan into the North Pacific (annual mean volume transport 1.5 Sv; 1 Sv =  $10^6 \text{ m}^3 \text{ s}^{-1}$ ) and is  
84 driven principally by sea level difference between the Sea of Japan and the North Pacific (e.g.,  
85 Ito et al., 2003; Toba et al., 1982). Primary productivity and nutrients increase downstream along  
86 the TWC (Matsuura et al., 2007; Saitoh et al., 2008; Yamada et al., 2005) and fishery resources  
87 (e.g., scallop, abalone, sea urchin, squid, and Pacific bluefin tuna) are abundant (e.g., Kosaka,  
88 2016; Sakurai et al., 2000; Shimose & Ishihara 2015). Recently, Ohta et al. (2015) revealed  
89 vigorous vertical turbulent mixing over the abrupt bottom of the sills in the western part of the  
90 Tsugaru Strait. Vertical turbulent flows mix the warm fresh upper water in the Tsugaru Strait  
91 with the cold saline deeper water transported by the Tsushima Warm Current in the Sea of Japan  
92 into the strait. The vigorous vertical turbulent mixing just east of the sills is thought to enhance  
93 nutrients and to maintain high productivity.

94 In the Sea of Japan (i.e., the upstream part of the TWC) subsurface waters at depths  
95 below 300 m are acidifying at a rapid rate ( $-0.002$  to  $-0.004 \text{ yr}^{-1}$ ), owing to the increasing  
96 atmospheric  $\text{CO}_2$  and increase in DIC associated with deoxygenation under reduced ventilation  
97 (Chen et al., 2017). The rapidly acidified water is thought to be transported into the Tsugaru  
98 Strait. Furthermore, the volume transport of the Tsushima Warm Current has increased  
99 considerably from 2005 to 2017, as observed at the eastern channel of the Tsushima Strait by  
100 ship-mounted Acoustic Doppler Current Profiler (Shibano et al., 2019). These changes in the Sea  
101 of Japan might enhance the ocean acidification in the eastern part of the Tsugaru Strait by  
102 vertical turbulent mixing in the western part.

103 To examine the extent to which ocean acidification has progressed at various depths in  
104 the Tsugaru Strait, we initiated a time-series observation of ocean acidification in February 2014  
105 by measuring temperature, salinity, nutrients, dissolved inorganic carbon (DIC), and total  
106 alkalinity (TA) by weekly bucket sampling at the breakwater of Sekinehama Port (BW) of the  
107 Mutsu Institute of Oceanography, Japan Agency for Marine-Earth Science and Technology  
108 (JAMSTEC), located on the northern coast of the Shimokita Peninsula of the eastern Tsugaru  
109 Strait (Figure 1). In addition, ship-based observations in the eastern part of the strait (stations  
110 SE3, SE9, OS1 and HO3) have been performed every season since the summer of 2012.



111  
112 Figure 1. Maps of the observation area and location of the sampling station near the islands of  
113 Japan. In the overview map (a), NSJ and OKH indicate the stations in the northern Sea of Japan  
114 and off Kushiro at Hokkaido island, respectively. Ryori is the time-series station for atmospheric  
115 CO<sub>2</sub> monitoring by the Japan Meteorological Agency (green open circle). The enlarged map (b)  
116 shows the time-series stations for ocean acidification monitoring in the eastern part of Tsugaru  
117 Strait (colored circles) and sea level monitoring (white open circles). Solid, dashed, and dotted  
118 lines indicate the 50, 100, and 200 m isobaths, respectively.

119

## 120 **2 Observations and Data**

### 121 2.1 Time-series observation of ocean acidification

122 We performed the time-series observations at the end of the breakwater (the water depth  
123 is 9 m) in Sekinehama Port (station BW) by bucket surface water sampling and conductivity-  
124 temperature-depth (CTD; RINKO-Profiler; JFE Advantech Co., Ltd.) measurements every week  
125 from February 2014 to December 2019 (Wakita, 2020). In this study, we used samples collected  
126 from 325 bucket casts at station BW.

127 The CTD SBE 911plus (Sea-Bird Scientific , Inc.) observations and water sampling with  
128 bucket and Niskin bottle were performed from the sea surface to the seafloor at stations SE3,  
129 SE9, OS1, and HO3 every season during 2010–2019 onboard Japanese training ships (*Ushio-*  
130 *Maru* and *Oshoro-Maru* of Hokkaido University: spring 2010 to fall 2019) and research vessels  
131 (*Wakataka-Maru* of the Japan Fisheries Research and Education Agency: summer 2018 and  
132 2019, *Kaiun-Maru* of the Aomori Prefectural Industrial Technology Research Center: winter  
133 2019). We started to collect water samples for DIC and TA from summer 2012. We used a total  
134 of 103 CTD casts in 35 cruises at stations SE3, SE9, OS1 and HO3.

135 As reference water properties for the water masses flowing into the Tsugaru Strait, we  
136 used physical and biogeochemical data from station NSJ (Sea of Japan Water) in the northern  
137 part of the Sea of Japan in 2013 (KS13-08) acquired by the R/V *Keifu-Maru* of the Japan  
138 Meteorological Agency (JMA) and station OKH (Coastal Oyashio Water) off Kushiro at  
139 Hokkaido island in 2015 (KH-15-1) acquired the JAMSTEC R/V *Hakuho-Maru*.

140 We used coulometric and potentiometric techniques to measure DIC with coulometers  
141 (CM5012, UIC, Inc; MODEL 3000A, Nippon ANS, Inc.) and TA using a total alkalinity titration  
142 analyzer (ATT-05, Kimoto Electric Co., Ltd.) (Wakita et al., 2017). The values were calibrated  
143 against certified reference material provided by Prof. A. G. Dickson (Scripps Institution of  
144 Oceanography, University of California San Diego) and reference material produced by KANSO  
145 Co., Ltd. The precision of the analyses for both the DIC and TA measurements was within  $\pm 0.8$   
146  $\mu\text{mol kg}^{-1}$ , based on replicate samples. The values of nutrients (silicate,  $\text{SiO}_2$ , and phosphate,  
147  $\text{PO}_4$ ) and salinity were measured with a continuous flow analyzer (QuAAtro, BL TEC K.K.) and  
148 a salinometer (Model 8400B AUTOSAL, Guildline Instruments Ltd.), respectively. These values  
149 were calibrated against certified reference material provided by KANSO Co., Ltd (for nutrients)  
150 and IAPSO Standard Seawater provided by Ocean Scientific International Ltd (for salinity).  
151 Based on replicate samples, the precisions for  $\text{SiO}_2$ ,  $\text{PO}_4$  and salinity were within  $\pm 0.2 \mu\text{mol}$   
152  $\text{kg}^{-1}$ ,  $\pm 0.01 \mu\text{mol kg}^{-1}$ , and  $\pm 0.001$ , respectively. The values of DIC, TA,  $\text{SiO}_2$ , and  $\text{PO}_4$  were  
153 normalized to a salinity of 35.0 to correct for the dilution and concentration of seawater by  
154 evaporation and precipitation and are expressed here as nDIC, nTA n $\text{SiO}_2$  and n $\text{PO}_4$ , respectively  
155 (e.g., nDIC = DIC  $\times$  35/salinity).

156 We used the CO2SYS program developed by Lewis and Wallace (1998) with the  
157 carbonate dissociation constants of Mehrbach et al. (1973) refitted by Dickson and Millero  
158 (1987), and values for temperature, salinity, DIC, TA,  $\text{SiO}_2$ , and  $\text{PO}_4$  to calculate the pH (in total  
159 hydrogen ion concentration scale) at the in situ temperature and  $25^\circ\text{C}$  ( $\text{pH}_T^{\text{in situ}}$  and  $\text{pH}_T^{25}$ ) and  $\Omega$   
160 ( $\Omega = [\text{Ca}^{2+}][\text{CO}_3^{2-}] \cdot K_{sp}^{-1}$  where  $K_{sp}$  is the apparent solubility product of calcite or aragonite;  
161 Mucci, 1983). Values of  $\Omega$  were calculated with respect to the mineral forms of  $\text{CaCO}_3$ ,

162 aragonite ( $\Omega_{\text{aragonite}}$ ). We estimated  $[\text{Ca}^{2+}]$  by assuming that  $[\text{Ca}^{2+}]$  was directly proportional to  
163 salinity (Millero, 1982).

164

## 165 2.2 Tide gauge data

166 The TWC is driven principally by sea level difference across the Tsugaru Strait (e.g.,  
167 Toba et al., 1982). Hourly sea level data are collected by the JMA at two coastal tide gauge  
168 stations, Fukaura and Hakodate (Figure 1b). Because the sea level difference between Fukaura  
169 and Hakodate is well correlated to the volume transport of the TWC, the sea level difference has  
170 been used as an indicator of volume transport of the TWC (e.g. Nishida et al., 2003). To correct  
171 the influence of vertical crustal movement, changes of the local reference levels observed on the  
172 first day of each year by the JMA (<https://www.data.jma.go.jp/kaiyou/db/tide/suisan/station.php>)  
173 were interpolated linearly at 1 hour intervals and subtracted from the hourly sea level at each  
174 station. In addition, the baselines of the tide gauge stations have been converted from the local  
175 reference levels to the mean sea level of Tokyo Bay. Further, tidal variations were suppressed by  
176 a tide-eliminating filter of Thompson (1983). After averaging to obtain daily data, atmospheric  
177 pressure correction was conducted using the daily mean sea level pressure observed at each tide  
178 gauge station. We used the daily mean sea level difference of Fukaura minus Hakodate.

179

## 180 2.3 Calculation of deseasonalized annual mean values

181 At the Tsugaru Strait, there are substantial seasonal variations in current velocity and  
182 water properties (e.g. Matsuura et al., 2007; Saitoh et al., 2008), which may yield a seasonal bias  
183 in the sampling and complicate the statistical trend analysis of time-series data (Bates et al.,

184 2014). Physical and biogeochemical data obtained by bucket surface water sampling are in  
185 agreement with those at 10 m depth from Niskin bottle water sampling (not shown). For the  
186 surface water down to 10 m depth, following Takahashi et al. (2006), we calculated mean values  
187 of the physical and chemical parameters for every month to obtain mean seasonal variations. For  
188 subsurface waters (below 20 m depth), we calculated mean values of the parameters for every  
189 season (winter from January to March, spring from April to June, summer from July to  
190 September, and fall from October to December) as in Bates et al. (2014), because the number of  
191 subsurface samples (103) is not sufficient to obtain significant mean values for each month.

192 For example, the deseasonalized monthly (seasonal) mean value of nDIC,  $nDIC_{\text{deseasonalized}}$ ,  
193 in the surface (subsurface) water was calculated as follows:

194

$$195 \quad nDIC_{\text{deseasonalized}} = nDIC_{\text{obs}}^{\text{month or season}} - nDIC_{\text{mean}}^{\text{month or season}} + nDIC_{\text{mean}}^{\text{annual}}$$

196

197 where  $nDIC_{\text{obs}}^{\text{month or season}}$  is the observed value of nDIC during a month for the surface water, or  
198 a season for the subsurface water.  $nDIC_{\text{mean}}^{\text{month or season}}$  is the surface monthly or subsurface  
199 seasonal mean value of nDIC, and sufficient  $nDIC_{\text{mean}}^{\text{annual}}$  is the annual mean value. The rates of  
200 change of these deseasonalized monthly and seasonal mean time series were calculated using a  
201 linear least-squares method.

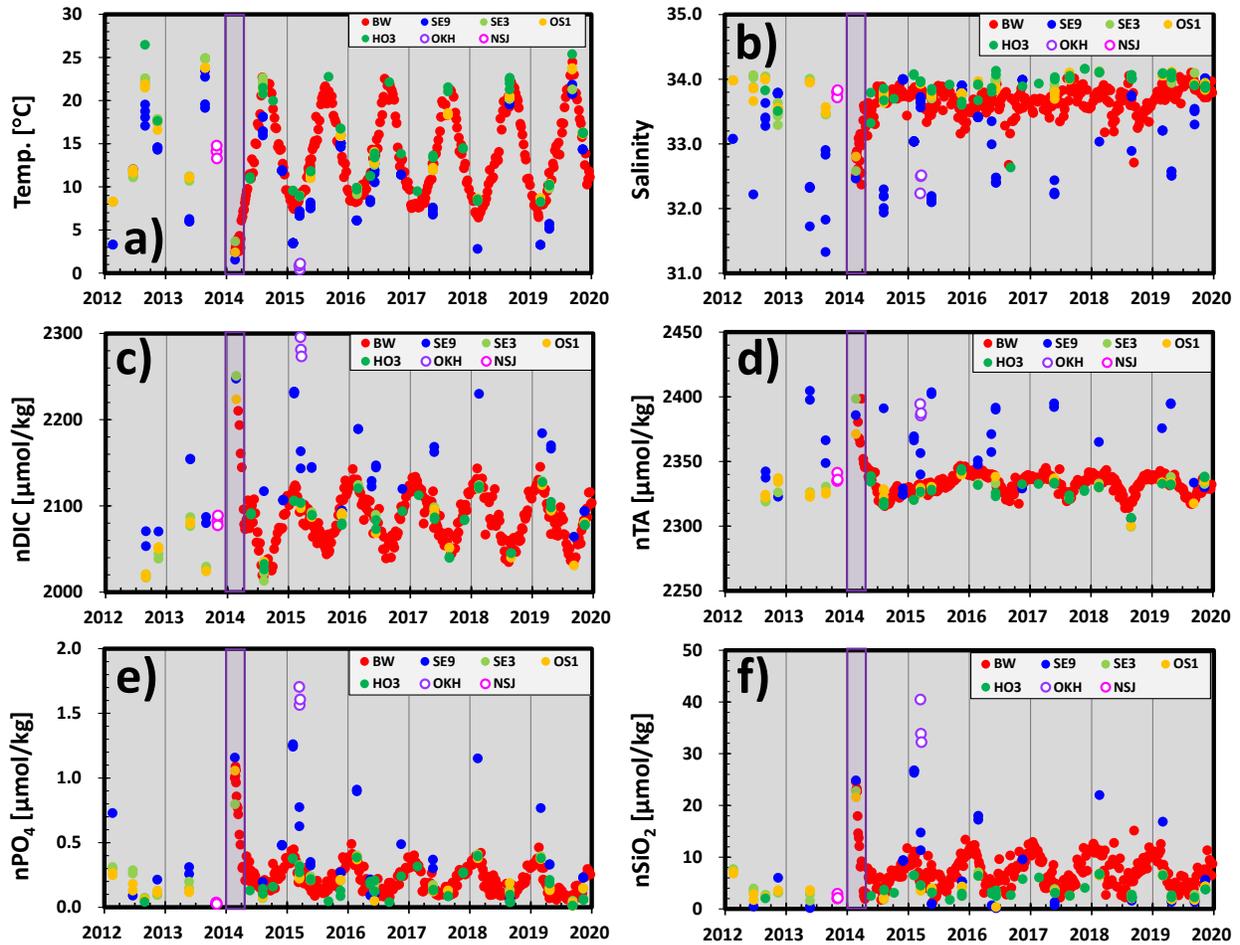
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### 203 **3 Results and Discussion**

#### 204 **3.1 Hydrographic features in the eastern part of the Tsugaru Strait**

205 At stations HO3, SE3 and OS1 in the eastern part of the strait, temperature, salinity,  
206 nDIC, nTA, and nutrients of the surface water have similar seasonal variations (Figure 2)  
207 because these stations were located in the surface water transported by the TWC throughout the  
208 year. However, the observed values during winter and spring at station SE9 are similar to those  
209 of the coastal Oyashio water (COW) at station OKH off Kushiro during the winter of 2015  
210 (purple circles in Figure 2). This similarity between the surface water properties at SE9 and OKH  
211 indicates that the surface COW was advected from the eastern coast of Hokkaido to the northern  
212 part of the Tsugaru Strait (e.g., Saitoh et al., 2008; Shimizu et al., 2001). Notably, values of the  
213 parameters observed at station BW in Sekinehama Port agree well with those at HO3, SE3, and  
214 OS1. In other words, the surface water obtained by bucket sampling at station BW has the  
215 characteristics of surface water of the TWC, except for during the period of COW advection  
216 from January to April in 2014 (purple hatches; Figure 2).

217 The seasonal variations of subsurface water properties at HO3, SE3 and OS1 are similar  
218 but differ from those at SE9 (Figure S1). The amplitudes of the seasonal variations decrease with  
219 depth. The advection of the COW from February to April in 2014 (purple hatches; Figure S1)  
220 extended to a depth of 20 m, as salinity, nTA and nDIC at SE9, SE3, OS1, and HO3 are similar  
221 to COW (purple circles). However, in this period, the water below 50 m depth at SE3, OS1, and  
222 HO3 was occupied by northern Sea of Japan water (SJW), as observed at station NSJ in the fall  
223 of 2013 (pink circles). Therefore, the eastern part of the Tsugaru Strait (stations HO3, SE3, OS,  
224 and BW) was dominated by TWC water through all depth layers from 2012 to 2019, except from  
225 during February to May in 2014 when it was affected by the COW and the northern SJW. In this  
226 study, we analyzed data at stations HO3, SE3, OS1, and BW to examine the temporal changes of  
227 TWC water after excluding the data collected during February to May in 2014.



228

229 Figure 2. Time-series of (a) temperature, (b) salinity, (c) nDIC, (d) nTA, (e) nPO<sub>4</sub>, and (f) nSiO<sub>2</sub>  
 230 in the surface water (0–10 m depth) in the eastern part of the Tsugaru Strait, in the northern part  
 231 of the Sea of Japan, and in the region off Kushiro, Hokkaido Island. Colors of plotted circles are  
 232 the same as those of the sampling stations shown in Figure 1. The period of the COW intrusion  
 233 from January to April in 2014 is indicated by purple hatches.

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### 238 3.2 Accelerated acidification of the TWC water

239 During 2012–2019, the deseasonalized annual mean values of  $\text{pH}_T^{\text{in situ}}$ ,  $\Omega_{\text{aragonite}}$ , and  
240 nDIC in the surface water and subsurface water changed considerably at rates of  $-0.0030$  to  
241  $-0.0051 \text{ yr}^{-1}$  ( $p < 0.05$ ),  $-0.017$  to  $-0.036 \text{ /yr}^{-1}$  ( $p < 0.05$ ), and  $1.9$  to  $3.6 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$  ( $p <$   
242  $0.05$ ), respectively (Figures 3 and 4), whereas nTA remained almost constant (Figures 2 and S1).  
243 These rates in the surface water ( $\text{pH}_T^{\text{in situ}}$ :  $-0.0030 \pm 0.0005 \text{ yr}^{-1}$ ;  $\Omega_{\text{aragonite}}$ :  $-0.017 \pm 0.003 \text{ yr}^{-1}$   
244 and nDIC:  $1.9 \pm 0.3 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$ ) were higher than those at time-series sites in other open  
245 oceans ( $\text{pH}_T^{\text{in situ}}$ :  $-0.0014$  to  $-0.0026 \text{ yr}^{-1}$ ;  $\Omega_{\text{aragonite}}$ :  $-0.002$  to  $-0.012 \text{ yr}^{-1}$  and nDIC:  $0.8$  to  $1.9$   
246  $\mu\text{mol kg}^{-1} \text{ yr}^{-1}$ ) summarized by Bates et al. (2014), and those predicted from oceanic  
247 equilibration with the increasing atmospheric  $\text{CO}_2$  ( $2.4 \mu\text{atm yr}^{-1}$ ) at station Ryori (Figure 1a)  
248 from 2012 to 2019 ( $\text{pH}_T^{\text{in situ}}$ :  $-0.0022 \text{ yr}^{-1}$ ;  $\Omega_{\text{aragonite}}$ :  $-0.0012 \text{ yr}^{-1}$  and nDIC:  $1.4 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$ )  
249 (Figure 4). The  $\text{pH}_T^{\text{in situ}}$  decline of the surface water is faster than that of the other coastal waters  
250 around Japan ( $\text{pH}_T^{\text{in situ}}$ :  $-0.002 \text{ yr}^{-1}$ ; Chen et al., 2017; Ishii et al., 2011; Ishizu et al., 2019; Lui  
251 & Chen, 2015).

252 In the subsurface layers, the rates of change of pH,  $\Omega_{\text{aragonite}}$  and nDIC were estimated to  
253 be  $-0.0043$  to  $-0.0051 \text{ yr}^{-1}$  ( $p < 0.05$ ),  $-0.024$  to  $-0.036 \text{ /yr}^{-1}$  ( $p < 0.05$ ), and  $2.8$  to  $3.6 \mu\text{mol}$   
254  $\text{kg}^{-1} \text{ yr}^{-1}$  ( $p < 0.05$ ), respectively (Figure 4). These rates are higher than for the surface water.  
255 This  $\text{pH}_T^{\text{in situ}}$  decline was faster than the highest rates ( $-0.004 \text{ yr}^{-1}$ ) in the North Pacific at  
256 200–500 m depth between 1991 and 2006 (Byrne et al., 2010) and in the Sea of Japan at 300 m  
257 depth between 1965 and 2015 (Chen et al., 2017). Evidently, the progress of acidification of the  
258 TWC water is rapid across all depths, particularly in the subsurface layer, compared with the  
259 surrounding oceans.

260 To identify the factors that account for the accelerated acidification, we examined the  
261 water properties that mainly control the rapid trends in  $\text{pH}_T^{\text{in situ}}$  and  $\Omega_{\text{aragonite}}$ . Note that  $\text{pH}_T^{\text{in situ}}$   
262 and  $\Omega_{\text{aragonite}}$  are functions of the pressure, temperature, salinity, DIC, TA, phosphate, and silicate  
263 of the TWC water. Following Wakita et al. (2013), we evaluated the trends in  $\text{pH}_T$  by allowing  
264 one parameter to vary while fixing the other parameters to the mean values. For example, we  
265 estimated the contribution of the nDIC or temperature to the decline of  $\text{pH}_T$  at 100 m depth by  
266 calculating the pH change using nDIC or temperature and mean values for the other parameters.  
267 The decline of  $\text{pH}_T^{\text{in situ}}$  ( $-0.0043 \text{ yr}^{-1}$ ) at 100 m depth is enhanced by nDIC ( $-0.0065 \text{ yr}^{-1}$ ) and is  
268 suppressed by temperature ( $+0.0022 \text{ yr}^{-1}$ ), because temperature in TWC during 2012–2019 did  
269 not increase significantly, despite global warming (not shown). The  $\text{pH}_T^{\text{in situ}}$  decline between 20  
270 m and 50 m ( $-0.0043$  to  $-0.0051 \text{ yr}^{-1}$ ) were slower than  $\text{pH}_T^{25}$  ( $-0.0058$  to  $-0.0075 \text{ yr}^{-1}$ ; Figure  
271 4). Namely, the temperature variation mitigates the  $\text{pH}_T$  decline. Thus, the increases in nDIC ( $1.9$   
272 to  $3.6 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$ ) are the main driver of the accelerated acidification, and are significantly  
273 higher than those due to the increasing atmospheric  $\text{CO}_2$  ( $1.4$  to  $2.0 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$ ; Figure 4).  
274 The rapid acidification is attributed to the enhanced rate of increase of nDIC caused by the  
275 oceanic uptake of anthropogenic  $\text{CO}_2$  and other physical and/or biogeochemical processes.

276 In addition to the trends of water properties, the volume transport of the TWC changed  
277 considerably during the study period. The sea level difference between Fukaura and Hakodate, a  
278 TWC transport index (e.g., Nishida et al., 2003), increased significantly at a rate of  $0.26 \pm 0.03$   
279  $\text{cm yr}^{-1}$  ( $p < 0.001$ ) during 2010–2019 (Figure 3f), consistent with the increase in volume  
280 transport of the Tsushima Warm Current reported by Shibano et al. (2019).

281 The intensification of the TWC is thought to increase the inflow to the Tsugaru Strait of  
282 deep water with high DIC associated with deoxygenation due to organic matter decomposition

283 under reduced ventilation in the Sea of Japan (Chen et al., 2017). This deeper water (saline, cold,  
284 high DIC) is possibly mixed into the upper layer via enhanced vertical turbulent mixing in the  
285 western part of the Tsugaru Strait,

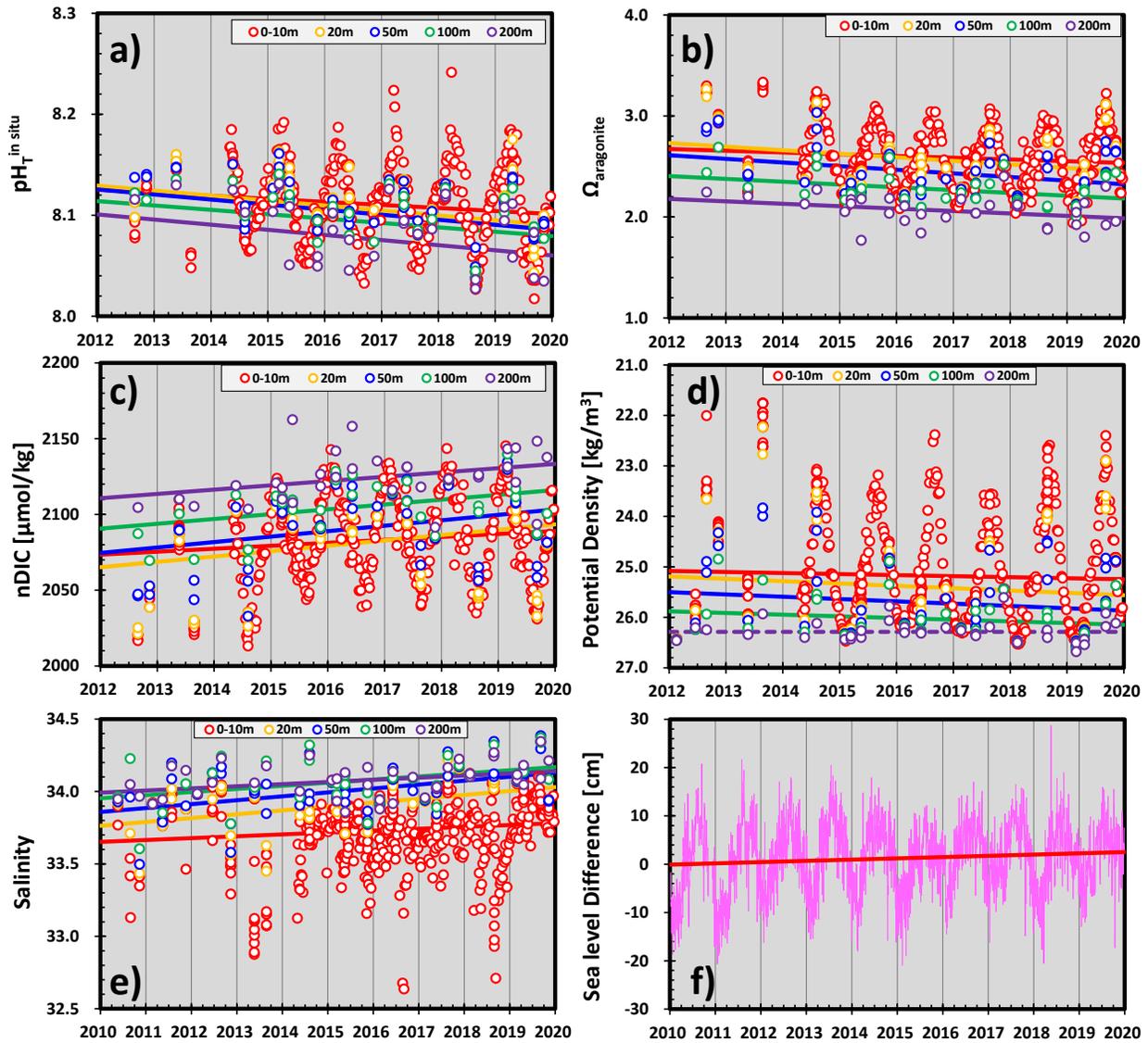
286 which is consistent with our observations of a significant increase in salinity, but not  
287 temperature, in the eastern part of the Tsugaru Strait (Figure 3e).

288 We discuss the impact of enhanced mixing of the deep high-DIC water in the Sea of  
289 Japan on the increase in nDIC in the eastern part of the Tsugaru Strait. Due to the significant  
290 salinity increase in the eastern part of the Tsugaru Strait, the potential density increased  
291 considerably (Figure 3d). The potential density is found to be linearly related to the nDIC  
292 increase (Figure S2). We use the linear relationship between potential density and nDIC to  
293 evaluate the impact of the increase of potential density water via enhanced mixing in the increase  
294 of nDIC. The slope of the Deming linear regression line (Deming, 1943) between nDIC and  
295 potential density is  $26.7 \pm 0.5 \mu\text{mol m}^3 \text{ kg}^{-2}$  ( $r = 0.89$ ,  $p < 0.0001$ ). Based on the regression, we  
296 calculated the rate of increase of nDIC due to the potential density increase ( $0.019$  to  $0.046 \text{ m}^3$   
297  $\text{kg}^{-1} \text{ yr}^{-1}$ ) to be  $0.5$  to  $1.2 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$ . The increase in nDIC expected from oceanic  
298 equilibration with increasing atmospheric  $\text{CO}_2$  is  $1.4$  to  $2.0 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$ . The sum of the two  
299 components is  $1.9$  to  $3.0 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$ , which is in good agreement with the observed nDIC  
300 increases ( $1.9$  to  $3.6 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$ ; Figure 4). Thus, the enhanced mixing of deep high-DIC  
301 water caused by the intensification of the TWC together with the absorption of atmospheric  $\text{CO}_2$   
302 are responsible for the enhancement of the increase in nDIC.

303 At 200 m depth, however, the calculated rate of nDIC increase ( $1.9 \pm 0.8 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$ )  
304 was slightly smaller than the observed rate ( $2.8 \pm 1.3 \mu\text{mol kg}^{-1} \text{ yr}^{-1}$ ). Because this occurred  
305 below the abrupt sill in the western part of the Tsugaru Strait ( $\sim 130$  m depth), the slightly smaller

306 increase rate of nDIC at 200 m depth might be caused by the downscaling of the enhanced  
307 mixing of deep high-DIC water or other processes, as there is no significant increases in salinity  
308 or potential density increase ( $p < 0.1$ ).

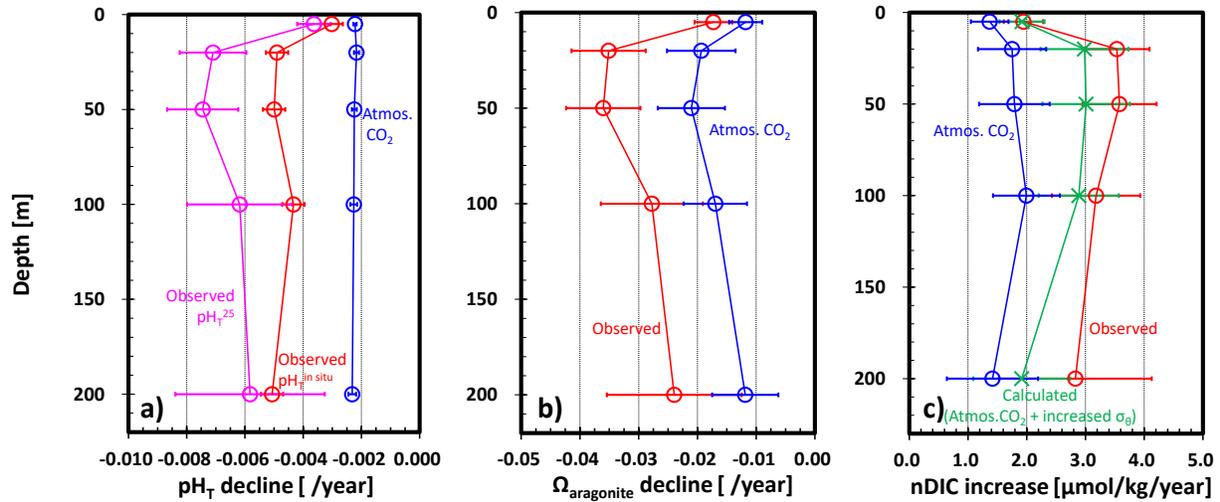
309



311

312 Figure 3. Time-series of (a)  $\text{pH}_T^{\text{in situ}}$ , (b)  $\Omega_{\text{aragonite}}$ , (c) nDIC, (d) potential density, and (e) salinity  
 313 in surface water (0–10 m depth, red circles) and subsurface (20 m depth, yellow circles; 50 m  
 314 depth, blue circles; 100 m depth, green circles; and 200 m depth, purple circles) water in the  
 315 eastern part of the Tsugaru Strait, and (f) sea level difference (pink curve) between Fukaura and  
 316 Hakodate (white circles in Figure 1). Solid (dashed) lines indicate statistically significant  
 317 (insignificant) regressions ( $p < 0.05$ ).

318



319  
 320 Figure 4. Rates of temporal changes in (a)  $\text{pH}_T$ , (b)  $\Omega_{\text{aragonite}}$ , and (c) nDIC as function of depth  
 321 during 2012–2019 in the eastern part of the Tsugaru Strait. Observed rates of  $\text{pH}_T^{\text{in situ}}$ ,  $\Omega_{\text{aragonite}}$ ,  
 322 and nDIC changes are indicated by red circles; predicted rates from the increasing atmospheric  
 323  $\text{CO}_2$  at a rate of  $2.4 \text{ ppm yr}^{-1}$  observed at the JMA Ryori atmospheric monitoring station (Figure  
 324 1a) during 2012–2019 by blue circles; the rates of  $\text{pH}_T^{25}$  decline by pink circles; and calculated  
 325 rates of nDIC increase due to the potential density ( $\sigma_\theta$ ) increase and the increasing atmospheric  
 326  $\text{CO}_2$  by green crosses. The error values represented by the horizontal lines are the standard errors  
 327 for the slope of the linear regressions.

328  
 329  
 330

### 331 4 Conclusions

332 We examined the acidification of the TWC water in the eastern part of the Tsugaru Strait  
 333 using time-series of physical and carbonate chemical data. Annual mean  $\text{pH}_T^{\text{in situ}}$  and  $\Omega_{\text{aragonite}}$  at  
 334 0–200 m depth in the TWC water decreased significantly at rates of  $0.0030\text{--}0.0051 \text{ yr}^{-1}$  and

335 0.017–0.036 yr<sup>-1</sup> during 2012–2019, respectively. Throughout all depths, the progress of the  
336 acidification at the Tsugaru Strait is found to be faster than that expected from the increasing  
337 atmospheric CO<sub>2</sub> and at time-series ocean sites in the open ocean and Japanese coastal waters.  
338 More noteworthy is the fact that the highest acidification is observed in the subsurface layer at  
339 20–100 m depth, and that is occurring faster than the highest rates observed during previous  
340 studies in the North Pacific Ocean and the Sea of Japan. The rapid acidification is found to be  
341 attributed to the enhanced increase of nDIC caused by the elevated mixing of deep high-DIC  
342 water into the upper layer by the intensification of the TWC in addition to the oceanic uptake of  
343 anthropogenic CO<sub>2</sub>.

344 The accelerated acidification observed at Tsugaru Strait may also occur under the  
345 influence of the TWC water in Mutsu Bay, Funka Bay, and the Sanriku coastal region (Figure 1),  
346 where there are the abundant fishery resources of calcifying organisms (e.g., scallop, abalone,  
347 and sea urchin). In other straits that are connected to the open ocean (e.g., the Tsushima and  
348 Soya Straits), the strengthening of their throughflow may also accelerate the acidification. At  
349 similar straits around the world, continuous monitoring of acidification is required to identify its  
350 progression and its effect on marine calcifying species and ecosystems.

351

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357 BW from 2014 to 2019 were downloaded from the Sekinehama Port carbon time-series data in  
358 the JAMSTEC Data Catalog  
359 ([http://www.godac.jamstec.go.jp/catalog/data\\_catalog/metadataDisp/Sekinehama\\_carbon\\_data?lang=en&view=simple](http://www.godac.jamstec.go.jp/catalog/data_catalog/metadataDisp/Sekinehama_carbon_data?lang=en&view=simple)). Physical and biogeochemical data were obtained in 2013 (KS13-08) by  
360 the JMA ([http://www.data.jma.go.jp/gmd/kaiyou/db/vessel\\_obs/data-](http://www.data.jma.go.jp/gmd/kaiyou/db/vessel_obs/data-)

362 report/html/ship/cruisedata\_e.php?id=KS1308). Atmospheric CO<sub>2</sub> contents at station Ryori were  
363 observed by the JMA from 2012 to2019  
364 ([https://www.data.jma.go.jp/ghg/kanshi/ghgp/co2\\_e.html#distribution](https://www.data.jma.go.jp/ghg/kanshi/ghgp/co2_e.html#distribution)). During 2010–2019, the  
365 JMA measured hourly sea level data and atmospheric pressure at two coastal tide gauge stations,  
366 Fukaura and Hakodate (<https://www.data.jma.go.jp/kaiyou/db/tide/genbo/index.php>;  
367 <https://www.data.jma.go.jp/gmd/risk/obsdl/index.php>; these data are not available in English).  
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380

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