

1 **Climate adaptation for a natural atoll island in the Maldives - predicting the long-**
2 **term morphological response of coral islands to sea level rise and the effect of hazard**
3 **mitigation strategies**

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9
10 **Key Points:**

- 11 • The natural island of Fiyooaree (Maldives) is projected to keep pace with 0.5m sea level
12 rise by 2100 through morphological adjustment.
- 13 • Ignoring the long-term morphological response of Fiyooaree to sea level rise triples the
14 predicted extreme overwash events by 2100.
- 15 • Reef adaptation measures reduce the inshore wave energy, limiting vertical accretion but
16 often mitigating the impact of extreme events.

17
18

19 Abstract

20 Coral atoll islands, common in tropical and subtropical oceans, consist of low-lying accumulations of carbonate
21 sediment produced by fringing coral reef systems and are of great socio-economic and ecological importance.
22 Previous studies have predicted that many coral atoll islands will become uninhabitable before the end of this
23 century due to sea level rise exacerbating wave-driven flooding. However, the assumption that such islands are
24 morphologically static, and will therefore 'drown' as sea levels rise, has been challenged by observations and
25 modelling that show the potential for overwashing and sediment deposition to maintain island freeboard. However,
26 for sustainable habitation, reliable predictions of island adjustment, flooding frequency and the influence of
27 adaptation measures are required. Here, we illustrate the effect of various adaptation measures on the morphological
28 response of an atoll island to future sea level rise using process-based model simulations. We found that the
29 assumption of a static island morphology leads to a significant increase in the predicted frequency of future island
30 flooding compared to morphodynamically active islands, and demonstrate that natural morphological adjustment is a
31 viable mechanism to increase island freeboard. Reef adaptation measures were shown to modify the inshore wave
32 energy, influencing the equilibrium island crest height and therefore the long-term morphological response of the
33 island, while beach restoration mainly delays the island's response. If embraced and implemented by local
34 communities, allowing for natural island dynamics and implementing well-designed adaptation measures could
35 potentially extend the habitability of atoll islands well beyond current projections.
36

37 Plain Language Summary

38 Low-lying coral atoll islands, protected by coral reefs, may become uninhabitable due to frequent flooding as sea
39 levels rise. However, studies also show that if left undisturbed by human activity, these islands could grow vertically
40 through the process of sand being deposited on the islands by waves washing over them. Understanding how these
41 islands will change and the impact of actions like coral reef restoration is crucial for planning sea level rise
42 adaptation. In this study, we used a computer model to see how an island in the Maldives will respond to different
43 sea level scenarios and adaptation measures. We found that if we assume that the island does not change over time,
44 we will predict far more flooding in the future than if we allow the island to adjust naturally to sea level rise. We
45 also found that some adaptation measures can reduce the damage from large waves that can lower and flood the
46 island, while further damage to the coral reefs can lead to more flooding. The results of the study show that allowing
47 the natural dynamics of coral atoll islands to take place may allow island communities to stay on their islands longer
48 than we thought.

49 1 Introduction

50 Atoll islands occur in the tropics and are wave-built accumulations of carbonate sediment derived from the
51 decomposition of calcium carbonate-secreting organisms that dwell on the adjacent, ring-shaped coral reef system.
52 The islands are of great socio-economic and ecological importance. They are the only naturally habitable land in
53 atoll nations such as Maldives, Tuvalu, Kiribati and the Marshall Islands, and in atoll island groups such as Chagos
54 and Lakshadweep. With a combined population of c. 1 M, atoll islands rely heavily on tourism for their GDP. They
55 also provide a critical habitat for diverse endemic and some threatened species (e.g., green turtles).
56

57 A characteristic feature of atoll islands is their low-lying nature (often < 1-2 m above high tide level), which makes
58 them vulnerable to coastal flooding and island inundation during extreme events (e.g., cyclones, storm surges,
59 tsunamis). Model-based projections of global mean sea level rise (SLR) by 2100 suggest an indicative range of
60 0.44–0.76 m SLR under the SSP2-4.5 ("middle of the road" Shared Socioeconomic Pathway) emission scenario
61 (*IPCC, 2021*). SLR will significantly increase the risk of flooding, particularly in the tropics, where even a 5-cm
62 SLR has been shown to double the exceedance probability of the 50-yr water level (*Vitousek et al., 2017*). It is
63 widely acknowledged that atoll islands are amongst the most vulnerable (coastal) environments to climate change
64 and it has been predicted that most of them will be uninhabitable by the mid-21st century due to SLR exacerbating
65 wave-driven flooding (*Storlazzi et al., 2018*).
66

67 Predictions of the impact of SLR on atoll islands are generally based on the application of process-based
68 hydrodynamic models (e.g., XBeach; *Roelvink et al., 2009*). Such models show overall good performance in
69 characterizing water levels and wave transformation over the reef, and correctly predict the increase in wave energy
70 along atoll island shores due to the elevated water depths on the reef platform (*Beetham and Kench, 2018; Jennath*

71 *et al., 2021*). These models can also account for the increase in flood risk due to reduced roughness of the reef
72 platform resulting from coral reef degradation (*Quataert et al., 2015*) due to local anthropogenic stressors
73 (overfishing, pollution) and global climate stressors (increase in ocean water temperature, ocean acidification). They
74 can also be used to assess the potential flood reduction from nature-based solutions such as coral restoration
75 (*Roelvink et al., 2021*).

76
77 Simulations of island flooding and inundation in response to sea level rise generally assume that the island itself
78 does not respond morphologically (i.e., by changing its shape or elevation). However, this is an oversimplification,
79 as atoll islands can adjust to rising sea levels by increasing their elevation as a result of waves overwashing the
80 island, transporting sediment from the reef platform and front of the island onto the top of the island. In fact,
81 geological studies argue that higher sea levels and associated high-energy wave events are actually critical for island
82 initiation and may even facilitate further vertical reef island building, provided there is an adequate sediment supply
83 (*East et al., 2018*). Depositional overwash processes have been observed *indirectly* on atoll islands through post-
84 storm surveys (*Smithers and Hoeke, 2014; Duvat and Pillet, 2017*) and *directly* through small-scale physical
85 modelling (*Tuck et al., 2019a; Tuck et al., 2019b; Tuck et al., 2021*). These experiments have been corroborated by
86 numerical modelling (*Masselink et al., 2020; 2021*) and demonstrate that repeated overwash occurring under high
87 wave and/or water level conditions can enable an atoll island to keep pace with SLR. Whether an atoll island can
88 maintain its freeboard (defined here as the difference between island crest elevation and mean sea level; *EurOtop,*
89 *2018*) during rising sea levels, is influenced by a number of factors, including the rate of SLR, storm wave climate
90 and sediment supply.

91
92 Allowing atoll islands to vertically accrete in response to SLR has significant implications for the assessment of
93 future coastal risk and results in less pessimistic predictions of future island habitability (*Masselink et al., 2020*). In
94 fact, provided there is sufficient sediment supply to the island and the flooding regime is not too frequent and
95 hazardous, island habitability is potentially sustainable. However, for communities to continue to occupy an atoll
96 island that is naturally adjusting to SLR through overwash-induced accretion demands flexibility and adaptability
97 from the current island infrastructure and communities, and a departure from the construction of hard-engineered
98 structures, as these prevent waves from overwashing and the island from morphodynamically adjusting (*Kench,*
99 *2012; Brown et al., 2020*). Constraints to living on such a naturally-adjusting island are also considerable and
100 include, amongst others, lack of space for the island to migrate to, resilience of island communities to episodic and
101 potentially hazardous flood events, adaptability of current island infrastructure and housing, threats to fresh water
102 supply and adverse impacts on agriculture. At the same time, early-warning systems (*Winter et al., 2020*) and
103 nature-based coastal protection measures (*Barnett et al., 2022*) can help facilitate island communities to sustainably
104 live on a dynamic and frequently-flooded island.

105
106 Adaptation measures, including nature-based solutions such as reef restoration and enhancing beach vegetation, may
107 help atoll islands to gradually adjust to SLR. However, quantifying the effectiveness of these measures requires
108 process-based, long-term morphodynamic modelling. Currently, an approach for such long-term process-based
109 morphodynamic modelling is lacking. XBeach is typically used to resolve complex hydrodynamic processes on
110 reefs (*Lashley et al., 2018; Quataert et al., 2020*) and the morphodynamic response during (a few) storm events
111 (*McCall et al., 2015; Masselink et al., 2020*), but is often considered computationally prohibitive for modelling
112 longer time scales. Conversely, wave-averaged models such as Delft3D are used to simulate long-term (longshore)
113 coastal evolution (*Luijendijk et al., 2019*), but lack second-order (low-frequency) wave dynamics, which tend to
114 dominate during overwash and flooding events.

115
116 Here, we use a small island (Fiyoaree) located on the exposed rim of Huvadho Atoll in the Maldives as a case
117 study to develop an efficient approach for investigating future atoll island habitability. Based on 40 years of
118 modelled wave data (ERA5) and measured water level data, we first generated an 80-year future synthetic wave and
119 water level time series. Wave events were filtered using an XBeach-based metamodel (BEWARE; *Pearson et al.,*
120 *2017*) to focus on only those events with significant expected morphological change. The cross-shore
121 morphodynamic development of the island was then modelled using the phase-resolving XBeach model. The
122 morphological response of the island and the changes in flood characteristics were compared for different adaptation
123 strategies.

124
125 Section 2 describes the study site, elaborates the six-step approach for developing future wave and water level time
126 series, and describes the numerical model setup for assessing the morphological response of the island. Section 3

127 then describes the wave transformation across natural and adapted reef profiles (Section 3.1), the morphological
128 response of natural coral islands (Section 3.2) and how adaptation measures influence the morphological response
129 (Section 3.3).

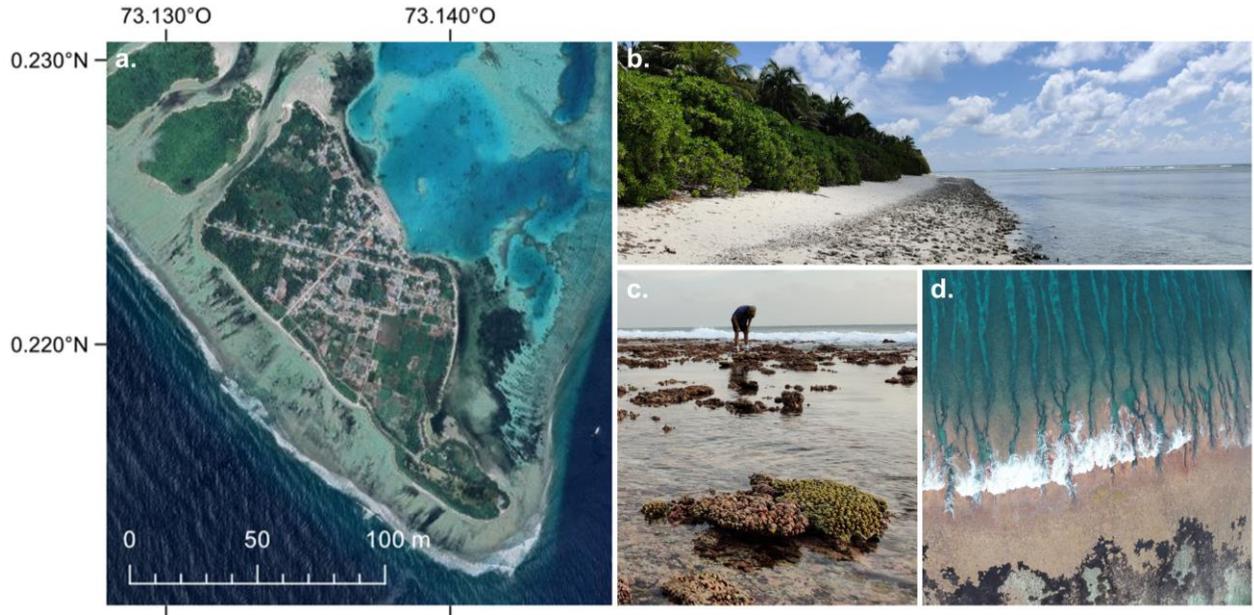
130 **2 Materials and Methods**

131 **2.1 Site description**

132 This study focuses on the populated island of Fiyoaree (c. 1,500 people) at the SSW rim of the Huvadho Atoll,
133 directly exposed to waves approaching from SE to NW. The island is approximately 1.6 km long and 0.6 km wide
134 and is fronted by a 100–200 m wide reef flat (see Figure 1a). The shallow reef flat is covered with alternating
135 patches of sea grass, sediment capturing sponges, and, among others, porites corals. The reef crest at the edge of the
136 reef flat is clearly defined and features a wide variety of coral types, such as acropora, heliopora, and platygyra. The
137 fore reef is composed of large spur (shore-normal ridge of coral) and groove (shore-normal channels) formations,
138 with maximum spur heights near the reef crest of up to 3 m above the adjacent groove and alongshore wavelengths
139 of 10–15 m (see Figure 1d). Like many islands on the southwestern rim of the Huvadho Atoll, Fiyoaree is
140 composed of a mixture of sand, gravel, and cobble-size material. Most of the ocean beach is lined with a
141 conglomerate platform composed of coral fragments, located around the mean sea level.

142
143 Water level variations due to tide and non-tidal residuals (NTR) are relatively low for the Maldives. The microtidal
144 regime (tidal range less than 2 m) is semidiurnal with strong diurnal inequalities. At Gan, the nearest tide gauge
145 located 100 km south of Fiyoaree, the mean spring tidal range is 0.96 m ([Wadey et al., 2017](#)). The NTR primarily
146 contains the meteorological contribution to sea level (surge). Wind-driven surge levels are low for Huvadho Atoll
147 since the atoll is located near the equator and therefore outside the normal range of cyclone activity. Furthermore,
148 wind-driven surge levels are generally low for the steep coastlines of coral islands. For steep coasts, and especially
149 outside the range of cyclones, the inverse barometer effect is generally the main contribution to total surge ([Van
150 Ormondt et al., 2021](#)).

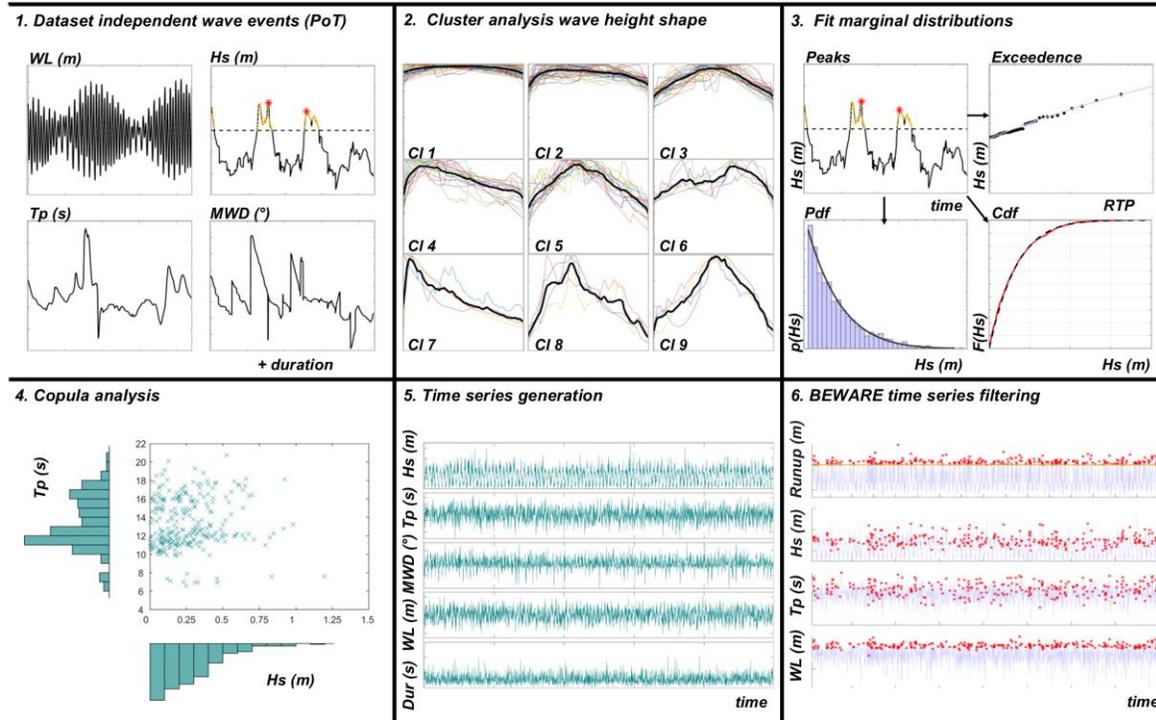
151
152 The wave climate of the Maldives is characterized by remotely generated swell waves, with largest significant wave
153 heights accompanied by peak periods of 10–12 s and up to 24 s ([Amores et al., 2021](#)). Persistent low-pressure
154 systems south of the Maldives (between 40 and 50 degrees south) generate year-round southwestern swells with
155 maximum significant wave heights of 4 m ([Amores et al., 2021](#)). During the southwest monsoon (May to
156 November), strong winds from the south to southwest generate long period and large swell waves ([Wadey et al.,
157 2017](#)), as well as shorter period sea waves. Fiyoaree is relatively sheltered from the northeasterly waves generated
158 during the northeast monsoon (November to April) but is fully exposed to the southwest swell.
159



160
 161 *Figure 1: Map of Fiyoaree Island on Huvadho Atoll (panel a) and photographs of the beach (panel b), reef crest (panel c), and*
 162 *spur and groove system (panel d).*

163 **2.2 Generating forcing conditions**

164 To assess how coral islands morphologically adjust to SLR (Section 2.3), we need a realistic future time series of
 165 wave conditions and water levels to force the morphodynamic model. While it is impossible to predict the exact
 166 sequence of future waves and water levels that will occur, the statistical characteristics of the past wave and water
 167 level climate can be used to generate multiple feasible realizations of the future time series. This time series can be
 168 used to force the XBeach model to generate a single feasible deterministic morphological prediction, or probabilistic
 169 predictions where computational effort allows. We describe the future wave and water level time series by a
 170 sequence of synthetic wave events, parametrized by the significant wave height H_s , the time evolution of H_s during
 171 the event, the peak wave period T_p , mean wave direction, storm surge and duration. A combination of advanced
 172 statistical techniques was used to generate reliable and realistic time series for the next 80 years (2020–2100). The
 173 method for deriving synthetic wave events involves six main steps (see Figure 2), as elaborated in the following
 174 sections.
 175



176
177 *Figure 2: Overview of the methodology to obtain an 80-year time series of wave and water level conditions to efficiently force a*
178 *morphodynamic model.*

179 **Step 1 – Obtain a dataset of independent wave events**

180 Water levels were obtained from water level measurements at the Gan tide gauge over the period 1992–2018. The
181 water level times series was detrended and was then raised to the current mean sea level to correct for sea level rise.
182 Non-tidal residuals are a function of various oceanographic and/or meteorological effects (e.g., wind, barometric
183 pressure, temperature, El Niño/La Niña effects) and were obtained by subtracting predicted tide levels from the
184 detrended water level measurements. Wind and wave climate variables were obtained from the ERA5 dataset
185 ([Hersbach et al., 2020](#)) from the European Centre for Medium-Range Weather Forecasts (ECMWF), which contains
186 hourly estimates of various atmospheric, land and oceanic climate variables. The wind and wave data are available
187 on a global grid with a resolution of 0.25 and 0.5 degrees, respectively, and cover the period from 1979 to the
188 present. ERA5 wave data have been extracted from the output node closest to Fiyooaree (0° N, 72.5° E). Table 1
189 shows the resulting monthly mean and 99 percentile wave heights for the four distinct wave seasons identified by
190 [LAMER \(2013\)](#).
191

192 *Table 1: Wave conditions at Fiyooaree summarized from the analysis of ERA5 wave data at coordinates 0° N, 72.5° E.*

| | Northeast (NE) monsoon | Transition Period 1 | Southwest (SW) Monsoon | Transition Period 2 |
|---|-----------------------------------|--------------------------------|-----------------------------------|--------------------------------|
| Period | December–February | March–April | May–September | October–November |
| Short wave direction | E-NE | NE-SE | S-SW | W |
| Swell wave direction | S-SW | S-SW | S-SW | S-SW |
| Mean significant wave height per month | 1.3 | 1.3–1.4 | 1.8–1.9 | 1.4–1.6 |
| 99 percentile significant wave | 1.7–2.0 | 1.8–2.2 | 2.5–2.6 | 2.4–2.5 |

height per month |

193

194 The peaks-over-threshold (PoT) method was used to identify individual wave events based on wave height peaks
 195 and a minimum duration between events. The wave time series was first divided into four distinct wave seasons (see

196 Table 1) to accurately represent the seasonal and monsoonal-driven variations in the wave climate of the Maldives.

197 The threshold for the POT method was defined as the mean plus one standard deviation of H_s , which equates to a
 198 threshold at the 68.3 percentile level, in line with [Masselink et al. \(2014\)](#). Note that although this threshold is

199 relatively low, it allows the inclusion of shorter return period events that may be relevant to the morphological

200 development of coral islands. Wave height peaks were identified using a minimum threshold of 12 hours between

201 peaks to ensure that selected peaks originate from independent events. Durations were defined as the time interval

202 associated with the wave height time series intersecting with the wave height threshold. This allows not only for the
 203 description of the peak H_s , peak T_p , and non-tidal residual water levels, but also of their evolution during an event.

204

205 **Step 2 – Identify the wave height shape from cluster analysis**

206 The events obtained in Step 1, each described by time series of wave conditions and the corresponding water levels,
 207 were parametrized for fitting marginal distributions (Step 3) and Copula (Step 4). Typically, wave events are

208 characterized by key parameters including the maximum wave height and the corresponding wave period and

209 direction during the event, as well as the duration. In addition to these parameters, we examined how wave heights

210 and periods change throughout an event. Consistent shapes were evident for H_s over the duration of the event, but

211 temporal variations in T_p were erratic, most likely reflecting a bimodal wave climate with peak periods alternating

212 between wind wave and swell periods. The most common shape of T_p was found to be almost constant and was

213 therefore adopted here. The evolution of H_s was parametrized by a single shape parameter using the k-means

214 algorithm ([Lloyd, 1982](#)), which partitions a data set into groups with similar characteristics. The number of wave

215 height shapes was optimized using the Silhouette index ([Rousseeuw, 1987](#)).

216

217 **Step 3 – Fit marginal distributions to the variables**

218 Optimized marginal distributions were fitted to the resulting dataset of variables to provide a statistical description

219 of the wave events, for each wave season separately. The marginal distributions were optimized using the AIC

220 criterion, a relative measure of how well a statistical model fits the data set from which it was generated ([Akaike,](#)

221 [1974](#)). The Generalized Pareto (GP), Generalized Extreme Value (GEV), Exponential, Weibull, and Rayleigh

222 distributions were evaluated for their goodness of fit. The significant wave height (Weibull), mean wave direction

223 (GEV), storm surge (Weibull), duration (Weibull), and wave height shape (Exponential) were well described by the

224 above distributions. Standard unimodal marginal distributions were less adequate to describe the bimodal or even

225 trimodal wave period shape but were chosen for their simplicity.

226

227 **Step 4 – Describe the interdependence between variables using Copula**

228 Copula analysis was used to establish the interdependence between pairs of variables for each wave season, linking

229 the wave height to each different variable. Copulas mathematically link random variables described by different

230 marginal distributions ([Nelsen, 2007](#)). Describing this dependency is important for accurately representing future

231 wave scenarios, as variables (e.g., wave height and wave period) can be strongly linked ([Poelhekke et al., 2016](#)). The

232 Frank copula family shows the best fit for most variable pairs, while the Gumbel copula was used to describe the

233 link between wave height and period.

234

235 **Step 5 – Generate synthetic wave events**

236 Synthetic wave events were generated by sampling the variables from the marginal distributions and copula

237 descriptions. The number of events in a given year and season was sampled from the historical number of events per

238 season. The events were then randomly distributed within each season to define the corresponding tidal water levels.

239 Ten synthetic wave event time series were generated, whose statistics fit well with the original data. A single time

240 series was randomly selected for further modelling efforts due to computational limitations. An SLR of 0.5 m by

241 2100 at a constant rate of 6.25 mm/yr was assumed, representing the median prediction for an intermediate

242 greenhouse gas (GHG) emission scenario (SSP2-4.5). The corresponding sea level rise values were added to the

243 offshore water level.

244

245 **Step 6 – Input reduction using BEWARE for efficient modelling**

246 For computational efficiency, the input conditions for the XBeach morphodynamic simulations (see Section 2.3)

247 were filtered using BEWARE2. BEWARE2 is an extension of the BEWARE ([Pearson et al., 2017](#)) wave runup

248 metamodel for coral reef environments that relates incident hydrodynamics and coral reef geomorphology to
 249 expected runoff. While the initial BEWARE database covers idealized fringing reef profiles, the BEWARE2
 250 database captures a range of real-world reef profiles defined based on the profile characterization by *Scott et al.*
 251 (2020). BEWARE2 was shown to accurately capture wave runoff at the Fiyooaree reef profile with an RMSE of 0.21
 252 m and bias of 0.05 m (see Supporting Information S1). Synthetic wave conditions with associated runoff below 1.75
 253 m, just below the island crest level, were excluded from the XBeach model simulations as they will not significantly
 254 influence vertical island adjustment.

255

256 2.3 Simulating the coral island's response using XBeach

257 The morphological response of Fiyooaree to the 80 years of synthetic wave events was simulated using XBeach, a
 258 process-based nearshore model that solves the depth-averaged, horizontal equations for flow, wave propagation and
 259 sediment transport (*Roelvink et al., 2009*). The XBeach-NH+ model was used, which solves the generation and
 260 transformation of both short and infragravity waves with an efficient reduced two-layer approach (*De Ridder et al.,*
 261 *2021*), enabling accurate predictions of waves and water levels across the reef and the resulting runoff and flooding
 262 (*Storlazzi et al., 2018*). The XBeach-G module within XBeach-NH+ enables the computation of groundwater flow
 263 using Darcy's Law modified for turbulent flow conditions and infiltration and exfiltration through the bed (*McCall*
 264 *et al., 2014*). XBeach-G also allows the computation of sediment transport and storm-induced bed level change for
 265 coarse sediments (bedload only) (*McCall et al., 2015*). Here we use the formulation of *Nielsen (2002)* for sediment
 266 transport on coral island beaches, as in *Masselink et al. (2020)*.

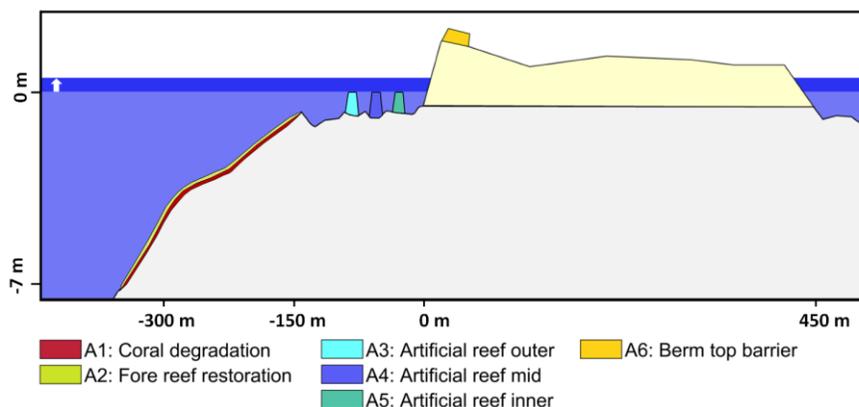
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268 A baseline one-dimensional XB-NH+ model was set up based on bathymetric measurements at Fares-Mathooda (a
 269 nearby island with similar characteristics to Fiyooaree for which bathymetric data were available), merged with the
 270 topographic profile of Fiyooaree. The grid resolution is high in shallow areas and on the beach to accurately simulate
 271 shallow water wave transformation and the island's morphological response. XB-NH+ parameters were adopted
 272 from *Roelvink et al. (2021)*. Reef friction values were parametrized following *Storlazzi et al. (2021)*, with friction
 273 values (c_f) of 0.05 for the reef flat, 0.1 for the fore reef and 0.003 for the beach.

274

275 The influence of five climate adaptation options was assessed by adapting the baseline XBeach model (see Figure 3
 276 and Table 2). The strategy of restoring reefs to improve coastal protection is increasingly proposed (*Viehman et al.,*
 277 *2023*) and quantified (*Brathwaite et al., 2022*). Therefore, four coral restorations were designed, including three
 278 artificial coral reef structures and an ecological fore reef restoration. Artificial coral reef structures were schematized
 279 as 10 m wide and 1.25 m high bed level elevations with increased roughness (c_f of 0.15), in line with *Storlazzi et al.*
 280 *(2021)*. Artificial reef locations were based on *Roelvink et al. (2021)*. Ecological fore reef restoration was modeled
 281 by increasing the roughness (c_f of 0.15) across the fore reef between depths of -7 m and -1 m, the lower limit set by
 282 the operational constraints specified by *Storlazzi et al. (2021)*. We also investigated the effect of strengthening the
 283 berm crest to reduce overwash and flooding events. The berm top barrier (*Lewis, 2021*) was simulated by adjusting
 284 the initial (sandy) island crest elevation (see Figure 3). In addition, simulations with reduced roughness (c_f of 0.01)
 285 across the fore reef and reef flat allow us to explore the influence of coral degradation.

286



287

288 *Figure 3: The coral island transect representative of Fiyooaree and the potential adaptation measures. The grey area shows the*
 289 *immovable coral island substrate, the light yellow the movable sandy island.*

290 *Table 2: Overview of the adaptation measures included in the modelling study.*

| Label | Name | Description |
|-------|-----------------------|--|
| A0 | No adaptation | Baseline scenario without adaptation measures. |
| A1 | Coral degradation | Reduced fore reef and reef flat roughness ($c_f = 0.01$) |
| A2 | Fore reef restoration | Increased roughness ($c_f = 0.15$) across the fore reef, between -7 and -1m depth. |
| A3 | Artificial reef outer | A 10 m wide, 1.25 m high artificial reef at the outer edge of the reef flat. |
| A4 | Artificial reef mid | A 10 m wide, 1.25 m high artificial reef at the middle of the reef flat. |
| A5 | Artificial reef inner | A 10 m wide, 1.25 m high artificial reef at the inner edge of the reef flat. |
| A6 | Berm top barrier | An increase in the height of the initial berm. |
| [-] | Static island | Model scenario without morphological bed updating. |

291
 292 Calibration tests were performed to define realistic settings for the sediment transport parameterization required for
 293 the morphological bed updating. The underlying assumption is that the current island is in equilibrium with the
 294 forcing, implying that it should not change significantly during regular storms and under the present sea level. The
 295 key tuning parameters are the hydraulic conductivity K and the Nielsen phase angle ϕ for sediment transport.
 296 Sensitivity tests were performed for K values of 0, 0.0005, 0.001, 0.005 m/s, and ϕ values of 28, 30, 32, and 34°. A
 297 K value of 0.0005 m/s combined with a ϕ value of 30° gave the most stable results for the first five years of wave
 298 events without sea level rise and without extreme overwash events, and was therefore adopted for the final
 299 simulations (see Supporting Figure S2).

300
 301 The 80-year time series consists of a sequence of synthetic wave events, each event described by its date of
 302 occurrence, duration, corresponding offshore water levels (the sum of the tidal water level, surge, and a sea level rise
 303 component), the wave height variation during the event, and a constant wave period (as detailed in Section 2.3).
 304 Each wave event was simulated individually, with the final bathymetric profile of one event serving as the input for
 305 the subsequent modelled event, assuming no morphological changes between separate events (waves not sufficiently
 306 energetic to reach the island crest). In total, 243 wave events were simulated, spanning a combined duration of 1114
 307 hours. XBeach required approximately 20 to 25 minutes per simulated hour on a single node of a basic laptop.

308
 309 The inundation patterns and the morphological evolution of the coral island were parametrized to assess the
 310 response of the island to sea level rise, considering different adaptation scenarios. For each wave event, the elevation
 311 and cross-shore position of the island's crest and the maximum overwash discharge were recorded. XBeach water
 312 level and velocity time series were post-processed in chunks of 1800 s to determine wave heights and water levels
 313 across the reef and the resulting runup. We used the method of *Guza et al. (1984)* to decompose the water level time
 314 series into incoming and outgoing components based on local water level elevations and current velocities. From
 315 spectral analysis, the corresponding total, incoming and outgoing wave height components were determined. A split
 316 frequency of 0.5 times the peak frequency of the incident waves was used to differentiate between sea-swell waves
 317 (denoted by SS; frequency greater than 0.5 times the peak frequency) and low-frequency waves (denoted by LF;
 318 frequency less than 0.5 times the peak frequency), as described by *Roelvink and Stive (1989)* and *Lashley et al.*
 319 *(2018)*.

320
 321 From the runup elevation time series, the $R_{2\%}$ runup levels were calculated, defined as the 2 % exceedance value of
 322 the runup peaks. To determine the different water-level contributions to the runup ($\eta_{2\%, \text{runup}}$) and incoming water
 323 level at the beach toe ($\eta_{2\%, \text{toe}}$), we also extracted the sea-swell (SS), infragravity (IG) and very low frequency (VLF)
 324 wave, and steady setup components from the runup and incoming water level elevation time series, respectively,
 325 following *Pearson et al. (2017)*. The steady setup component was calculated as the mean water level relative to the
 326 still offshore water level, while the SS, IG, and VLF wave components were obtained from the detrended runup and
 327 beach toe water level time series by spectral analysis. The total water level (relative to mean sea level) and its
 328 components were then sorted in ascending order to select the 2% exceedance value.

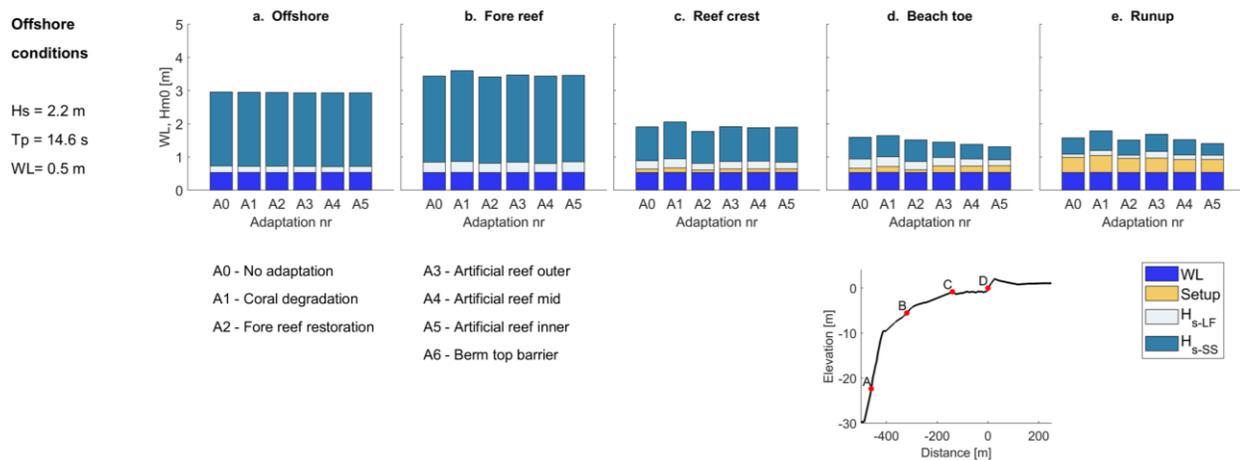
329

330 **3 Results**

331

332 **3.1 Wave transformation across natural and adapted reef profiles**

333 Figure 4 shows the wave transformation across the natural Fiyooree coral reef profile during an energetic synthetic
 334 wave event, the third wave event in the 243-event time series. Wave transformation across this profile is
 335 characterized by shoaling across the fore reef (panels a and b in Figure 4), wave breaking across the upper fore reef
 336 and reef crest (panels b and c in Figure 4) and subsequent dissipation across the reef flat (panels c and d in Figure 4).
 337 Incident wave reflection is limited at the gently sloping (1:20) upper fore reef, while reflection of low-frequency
 338 (LF) waves (VLF and IG components) at both the fore reef and beach can be significant. Wave breaking in the
 339 relatively narrow surf zone induces a significant increase of the mean water level across the reef (setup). The time-
 340 varying location of the breakpoint is an important driver of the nearshore IG wave energy, as also observed by
 341 [Merrifield et al. \(2014\)](#) and [Becker et al. \(2016\)](#), among others. On the inner reef flat, very-low frequency (VLF)
 342 energy peaks were observed when the LF peak frequency at the reef crest coincided with the natural resonance
 343 frequency of this reef, which is a function of the length and depth of the reef flat ([Péquignot et al., 2009](#)). The VLF
 344 energy peak around the fundamental resonant mode (see Supporting Figure S3) is an indication of the occurrence of
 345 resonance on this relatively narrow and smooth reef flat. A second surfzone is present near the island beach, where
 346 radiation stress forced set-down and setup can be clearly identified (not shown here).
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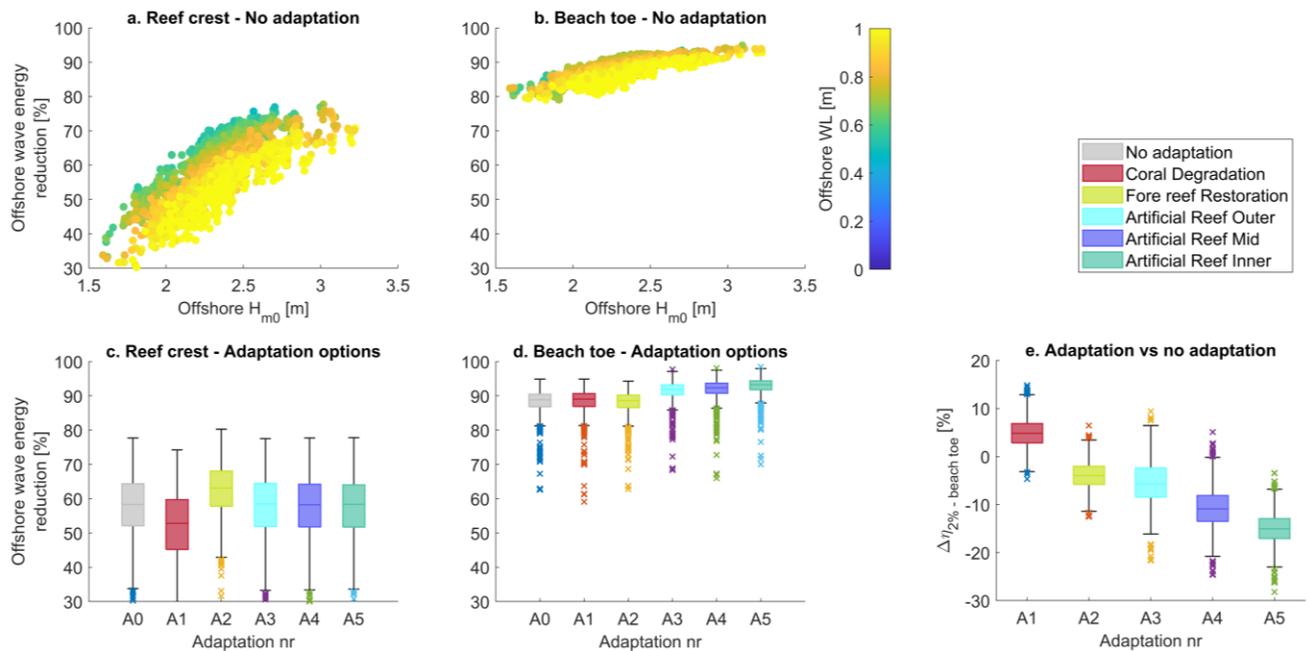


349
 350 *Figure 4: Wave transformation across both unrestored (A0) and adapted (A1 to A5) reef morphologies, depicting water level and*
 351 *total wave height components at four locations across the reef (columns; a. offshore, b. fore reef, c. reef crest, and d. beach toe)*
 352 *as well as the $\eta_{2\%,runup}$ (panel e) for a single synthetic wave event with a significant wave height of 2.2 m, peak wave period of*
 353 *14.6 s and offshore water level of 0.5 m.*

354 Wave transformation and setup across a reef, and the resulting runup, are highly dependent on the incident wave
 355 conditions and water levels, as stipulated by [Becker et al. \(2014\)](#) and [Pearson et al. \(2017\)](#). Here, we investigated
 356 the influence of extrinsic conditions on the incoming water level components at the beach toe (setup, VLF, IG, and
 357 SS; Supporting Figures S4-S7) instead of the runup, as the runup levels are (more) influenced by the morphological
 358 response of the island and are curtailed by the island crest. The setup across the reef shows a strong correlation with
 359 the offshore wave power (see Supporting Figure S4), while the modelled water level variations were too small to
 360 show significant trends. VLF waves were only observed for long period incident wave conditions (T_p longer than 18
 361 s, see Supporting Figure S5) and for reef flat water depths larger than approximately 1.4 m, during which resonant
 362 modes could be excited. The incident wave period has a significant control on the incoming IG wave heights, mainly
 363 through reflection at the fore reef. The long period incident waves generated long period LF waves at greater water
 364 depths (through intensified shoaling), in this case the steeper part of the forereef, which was likely to increase their
 365 reflection. As a result, the incoming IG wave height at the reef flat was higher for shorter period waves (see
 366 Supporting Figure S6), although the total (incoming plus outgoing) IG wave energy is higher for longer period
 367 waves due to the large reflection at the beach. The nearshore short wave energy was mainly dictated by reef flat
 368 water depths (Supporting Figure S7). Combining the wave attenuation statistics across the range of model variations
 369 during the 243 wave events, the incident wave energy is reduced by 30 to 80% at the reef crest of the unrestored reef
 370 profile (median value of 58 %; Figure 5a,c), with the percentage of energy reduction increasing with the offshore
 371 wave height (see Figure 5a). Only 5 to 20 % of the offshore wave energy reaches the shore (median value of 11 %;
 372 Figure 5b,d), and this value decreases with increasing offshore wave height (Figure 5b).
 373

374 Adaptation strategies influence wave transformation by promoting wave energy dissipation through friction and
 375 wave breaking. The fore reef restoration (A2) reduces the setup (especially for low wave power events, see
 376 Supporting Figure S4) and infragravity (IG) wave heights (especially those generated during low wave period
 377 events, see Supporting Figure S6). In contrast, coral degradation (A1) decreases the dissipation of incident wave
 378 energy over the forereef, allowing more wave energy to be transmitted to the reef flat. The increase in radiation
 379 stress gradients as a result of reduced short wave dissipation amplifies the setup and breakpoint-generated low-
 380 frequency waves (Figure 4c, Figure 5c). Enhanced frictional dissipation across the fore reef reduces extreme beach
 381 toe water levels by 4 % (median; see Figure 5e) to a maximum of 12 %, while coral degradation increases extreme
 382 water levels by 5 % (median; Figure 5e).

383
 384 The effectiveness of artificial reefs (A3, A4, and A5) in reducing coastal hazards is strongly dependent on their
 385 location and the prevailing forcing. Located close to the wave breaking point, the outer artificial reef structure (A3)
 386 can increase rather than decrease runup (see Figure 5e) due to the amplification of the setup and IG motions caused
 387 by the enhanced radiation stress gradients (see Supporting Figures S4 and S6), as also observed by *Roelvink et al.*
 388 (2021). Additional setup by the artificial reef (likely overestimated in a 1D model, see Discussion) is mainly
 389 observed during low water levels and relatively low offshore wave power. The increased water level across the reef
 390 flat in turn reduces depth-induced wave breaking and frictional wave dissipation, further exacerbating the risk of
 391 flooding. The artificial reef structures closer to shore (A4 and A5) locally increase the setup but generally reduce the
 392 wave-driven contribution to runup, especially by reducing the short wave energy reaching the shore (Supporting
 393 Figure S7). Overall, the median extreme water level reductions are 6, 11, and 15% for the outer, middle, and inner
 394 artificial reef structures, respectively, with maximum reductions of almost 30% (see Figure 5e).
 395



396
 397 *Figure 5: Offshore wave energy reduction at the reef crest (panel a) and beach toe (panel b) of the unrestored reef profile (A0),*
 398 *colored based on the offshore water level. Panels c and d depict offshore wave energy reduction values at the reef crest and*
 399 *beach toe for both unrestored (A0) and adapted (A1 to A5) reef morphologies. The reduction in extreme beach toe water levels*
 400 *($\Delta\eta_{2\%}$; positive values denote an increase in extreme water levels) of adapted reefs relative to the unrestored reef is depicted in*
 401 *panel e. The boxplots in panels c, d, and e depict the median, lower (25%), and upper (75%) quartiles (colored boxes), the*
 402 *minimum and maximum values not considered outliers (black lines), and the outliers (crosses).*

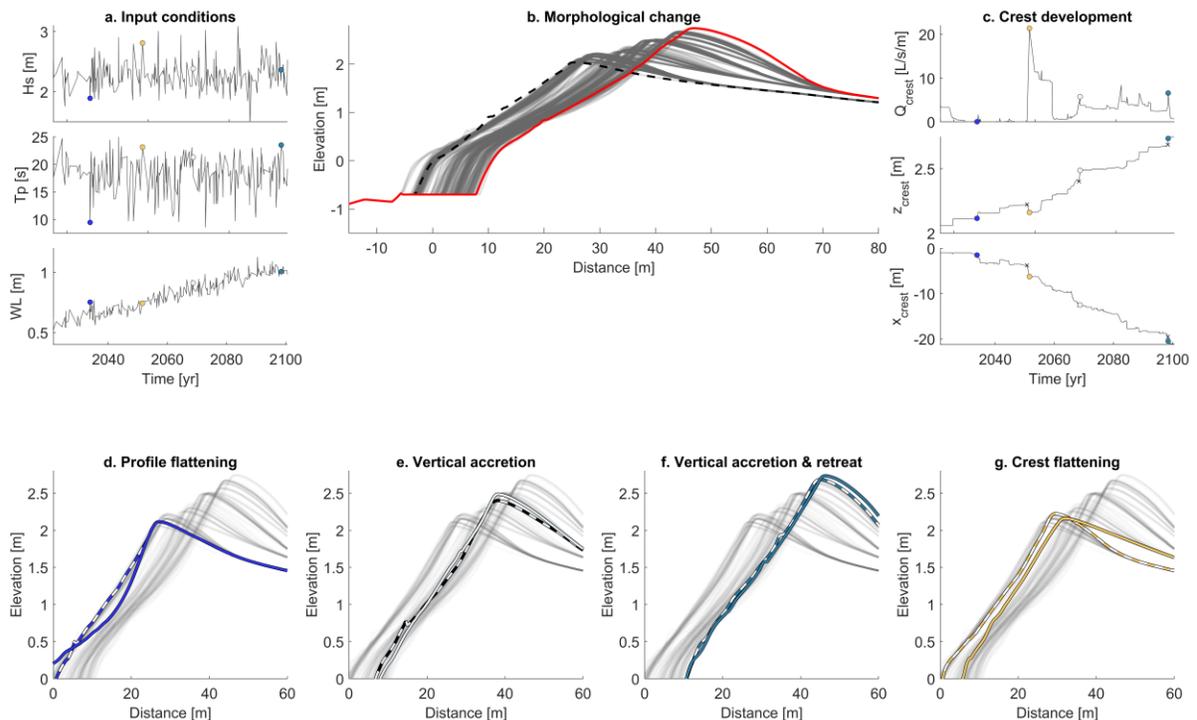
403 3.2 Morphological response of natural coral islands to SLR

404 The modelled response of the natural coral island beach profile over the 80-year simulation period shows significant
 405 morphological change, including profile flattening and steepening (Figure 6d), vertical accretion (Figure 6e), island
 406 retreat (Figure 6f) and crest flattening (Figure 6g). Low-energy, short-period wave events tend to flatten the lower

407 beach profile and steepen the upper beach profile, creating a wide berm as observed by [Jackson et al. \(2002\)](#) (see
 408 Figure 6d). The inflection point between the steeper upper beach face and lower-gradient lower beach face is
 409 strongly controlled by the water level. High-energy, longer-period wave events re-steepen the beach profile to the
 410 approximate shape observed in Figure 6e–g. Limited overwash leads to vertical accretion of up to 0.1 m during a
 411 single wave event (Figure 6e), whereas there is retreat of up to 1 m during more intense wave events (Figure 6f).
 412 Figure 6c reveals the sensitive threshold between vertical accretion with and without island retreat, where a slight
 413 increase in overwash discharge induces crest retreat (compare panels e and f of Figure 6). Extreme wave events with
 414 mean overwash discharge exceeding the 10 L/m/s threshold observed by [Masselink et al. \(2020\)](#) lead to a lowering
 415 (-0.06 m) and retreat (2.5 m) of the island's crest (see Figure 6c and 6g).

416
 417 As sea level rises, the island's crest gradually builds up and retreats over time (Figure 6b), as also demonstrated by
 418 [Masselink et al. \(2020\)](#). During the first 10–15 years of wave events, the morphological adjustment of the island is
 419 limited, suggesting that the current island morphology is not in equilibrium with the forcing. It is worth noting that
 420 there is strong evidence that atoll islands in the Maldives were established during the mid-Holocene with slightly
 421 higher sea level ([Kench et al., 2005](#)). The implication is that these islands are somewhat 'attuned' to a sea level
 422 elevation above current sea level and this, perhaps, explains the delay in island response to SLR in the simulation.
 423 Another explanation is a slight disequilibrium in the model, and that wave runup may in reality be slightly higher
 424 than predicted, i.e., matching the current beach crest. From 2035 onwards, the island crest accretes vertically not
 425 only due to the rise in nearshore water levels, but also due to the concurrent increase in wave energy (when reef flat
 426 accumulation does not keep pace with SLR). Periods following destructive events such as a hypothetical event in
 427 2051 (Figure 6g) are followed by accelerated crest accretion (Figure 6c), while periods following highly accretive
 428 events (Figure 6f) are followed by more mild accretion. After the high energy wave event in 2051, retreat rates are
 429 relatively constant. Over the 80-year period, the island has accreted by 0.7 m, keeping pace with the 0.5 m SLR and
 430 the concurrent increase in nearshore wave energy. However, the threshold between crest accretion and crest
 431 lowering and retreat, or even complete island rollover, is likely sensitive and could shift with higher rates of SLR or
 432 changes in storm intensity and frequency. For Fiyoaree, the assumption of a static island morphology that does not
 433 evolve with SLR (compare black and grey lines in Figure 7) leads to an increase in overwash hours by 2100 by
 434 almost a factor of three (Figure 7c). Thus, ignoring the natural morphological response of the island to sea level rise
 435 when projecting future flood risk is likely to result in an overly bleak outlook for the habitability of this island.

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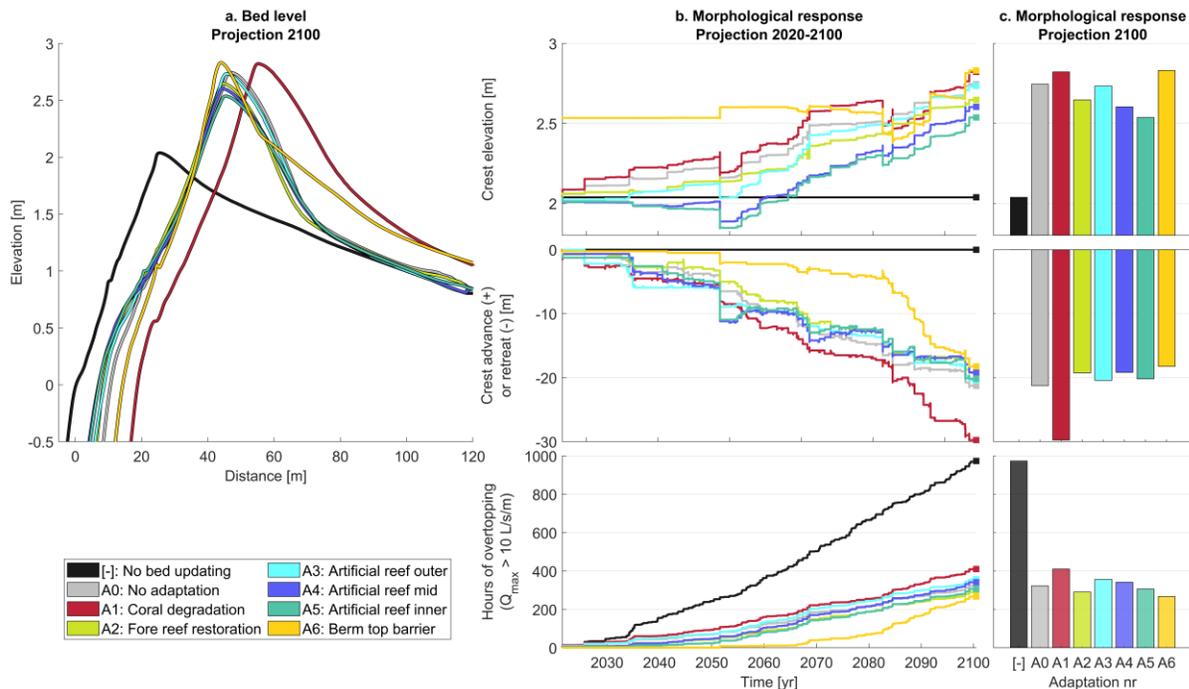


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439 *Figure 6: Morphological change of the Fiyooaree reef profile without interventions, depicting the input conditions (panel a), the*
 440 *resulting morphological change over time (panel b), and the development of the beach crest in terms of crest discharge, crest*
 441 *elevation, and cross-shore position of the crest (panel c). Panels d–g indicate different modes of the morphological response, for*
 442 *which the corresponding input conditions and crest development indicators are color-coded in panels a and c. Dashed, grey, and*
 443 *colored lines show the initial, intermediate, and final profiles, respectively.*

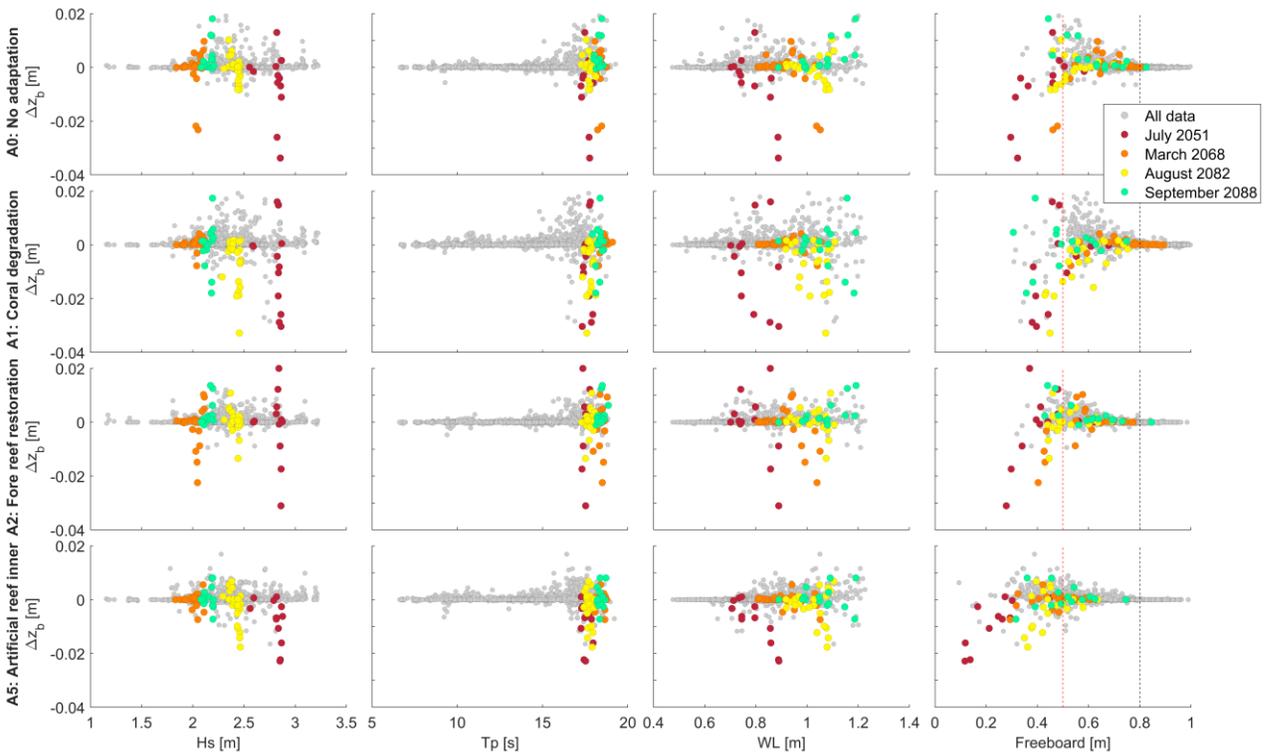
444 3.3 Morphological response of reef profiles with adaptation

445 Figure 7 shows the projected bed levels by 2100 for the natural and adapted island simulations (grey and colored lines) and the static bed (black line) (Figure 7a), the morphological response over time in terms of crest elevation,
 446 crest retreat and hours of overwash (Figure 7b), and the projection of the morphological response and inundation for
 447 2100 (Figure 7c). Adaptation measures either modify the nearshore wave energy (A1–A5) or limit overtopping (A6),
 448 influencing accretion and retreat rates, thereby ultimately shaping the island's morphology. Enhanced nearshore
 449 wave energy and wave setup due to coral degradation (A1) leads to large vertical accretion ($dz = +0.8$ m), but also a
 450 higher frequency of overtopping (410 hours of overwash exceeding 10 L/m/s) and retreat ($dx = -30$ m) compared to
 451 the natural profile, which experiences an accretion of +0.7 m, 320 hours of overwash, and a retreat of 21 m. In
 452 contrast, fore reef restoration reduces the overtopping frequency by 6% (290 hours of overwash exceeding 10 L/m/s)
 453 and retreat by 10% ($dx = -19$ m) compared to the natural profile, with less vertical accretion ($dz = +0.65$ m). The
 454 inner artificial reef, which is most efficient in wave height reduction, exhibits the lowest increase in crest height
 455 after 80 years of wave events ($dz = +0.5$ m), with a frequency of intense overwash (305 hours) similar to the natural
 456 profile. Given that the outer and mid artificial reef structures can amplify setup and LF motions, they do not reduce
 457 the total number of hours of more intense overwash events (Figure 7c, bottom panel). Nonetheless, they do result in
 458 lower retreat and crest accretion values. Evidently, the island crest height is in equilibrium with the nearshore wave
 459 energy level; therefore, the more energy dissipation across the fore reef and reef flat, the lower the crest elevation.
 460 The berm top barrier shows a highly variable behavior over the years. By 2065 there is hardly any overtopping and
 461 hence limited accretion ($dz = +0.05$ m) or retreat ($dx = -4$ m). The island crest has become steep and narrow in
 462 contrast to other island profiles that have accreted vertically, but that have also widened during periods of overwash.
 463 Consequently, several events around 2080 contribute to a significant crest lowering ($dz = -0.2$ m) and a large retreat
 464 ($dx = -12$ m) is observed between 2080 and 2100. The rate of retreat is strongly dependent on the initial design of
 465 the berm top barrier. Wide(r) berm enhancement works are likely to mimic the natural response of the island to SLR
 466 without flooding or island retreat.
 467
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 469



471 *Figure 7: Modelled bed level projections for 2100 (panel a), the morphological response indicators over time (panel b), and the*
 472 *morphological response projection for 2100 (panel c) for the unrestored (A0) and adapted (A1 to A6) reef morphologies, as well*
 473 *as the static bed scenario ([-]). The morphological response is quantified by the crest elevation, crest retreat, and cumulative*
 474 *hours of overtopping.*

475 Figure 8 shows the morphological response of the island, parametrized by the crest elevation change Δz_b , plotted
 476 against the extrinsic conditions (offshore wave height, wave period and water level) as well as the island freeboard
 477 (here defined as the difference between the island crest height and the extreme water level at the beach toe $\eta_{2\%, \text{beach}}$),
 478 for a subset of the adaptation scenarios. Two hypothetical extreme events resulted in significant reductions in crest
 479 height and retreat in the model: the 2051 event (see also Figure 6g and Figure 7) and the 2082 event (see also Figure
 480 7). During the 2051 event, inundation primarily stems from extreme wave conditions (large wave height, long wave
 481 period), whereas the 2082 event is characterized by long-period waves over elevated reef flat water levels. Degraded
 482 reefs are particularly vulnerable to the 2051 and 2082 wave events, with a total reduction in crest height (sum of
 483 markers per event in Figure 8) of 0.1 m and 0.15 m respectively, in contrast to total reductions of 0.06 m and 0.01 m
 484 for the natural reef, and almost negligible change for the restored reef. Although the artificial reef structures do not
 485 perform well under the energetic wave conditions of 2051 due to the low water levels (resulting in 0.08 to 0.12 m of
 486 crest lowering), they cause only modest crest lowering in 2082. Degraded reefs experience wave events that cause
 487 crest lowering more frequently than other adaptation scenarios. Yet, there are also more instances of crest accretion,
 488 allowing vertical growth in response to SLR and the increased inshore wave energy. The difference between the
 489 island crest height and extreme beach toe water levels, indicated by the freeboard in the right panels of Figure 8,
 490 provides a first indication of the morphological response of the island. When the freeboard exceeds 0.8 m, no
 491 changes occur. Within the range of 0.5 and 0.8-m freeboard, there is a tendency for crest accretion. Below 0.5-m
 492 freeboard, the island crest generally tends to lower. The subtle variations in forcing between the 2068 and 2088
 493 events with divergent morphodynamic response highlight the sensitivity of the threshold for crest elevation change.
 494



495 *Figure 8: The crest elevation change (y-axis) at half-hourly time intervals for all wave events (grey markers) and four specific*
 496 *events (colored markers), plotted against the offshore wave height (H_s), wave period (T_p), and water level (WL), as well as the*
 497 *freeboard, which was defined as the difference between the island crest height and the extreme beach toe water level ($\eta_{2\%, \text{beach}}$).*
 498 *The rows depict results for four different scenarios; A0: No adaptation, A1: Coral degradation, A2: Fore reef restoration, and*
 499 *A5: Inner artificial reef, chosen for their distinct morphological behavior.*
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Discussion

Atoll islands are dynamic landforms that evolve under the influence of wind, waves, currents, and sea level (Kench *et al.*, 2009) with the potential to adapt and remain habitable through overwash-induced accretion. This is true for more pristine coral island types, while on heavily urbanized islands (such as Malé), engineering solutions are often favored to maintain their functionality (Kench *et al.*, 2011). In this study, we used the XBeach-NH+ model to simulate the morphological response of Fiyoaree, a sparsely populated island in the Maldives, to 80 years of synthetic wave events under SLR. Various adaptation measures, including fore reef restoration, artificial reefs on the reef flat, coral degradation, and berm restoration were examined for their influence on reef hydrodynamics and the island's morphological response. Results revealed that artificial reefs can reduce nearshore extreme water levels by 15% compared to the 'no-adaptation scenario', with a maximum reduction of approximately 30%. Enhanced frictional dissipation from fore reef restoration reduced nearshore extreme water levels by 4%, while coral degradation resulted in a 5% increase. A clear balance was observed between the wave energy reaching the toe of the island and the elevation of the island crest, with the lowest equilibrium crest heights observed for the adaptation measures most effective in attenuating wave energy. As a result, the island crest height of adapted reefs initially exceeds its equilibrium value, reducing the frequency of overwash events. However, as sea levels continue to rise, adapted islands respond morphologically through accretion and retreat. Consequently, once the islands are in equilibrium with their overwash regime (a few decades into most simulations), fairly similar future flood frequencies are projected between adapted and unrestored islands. However, adaptation measures, especially fore reef restoration, can reduce the impact of the more extreme events that lead to significant island crest lowering and retreat. Degraded reefs are particularly vulnerable to these events, showing highest levels of crest lowering and retreat for specific events. The berm top barrier provides protection with limited nuisance flooding or retreat until 2080. Its narrow design strongly influences the progressive retreat after 2080, which wider and more regular berm enhancement works may prevent.

According to the model simulations, overwash events that resulted in crest accretion were characterized by overwash discharges of less than 10 L/m/s. Higher discharges resulted in the lowering of the island crest and accretion of the island landward of the crest, agreeing with findings from previous studies (Masselink *et al.*, 2020; 2021). Another indicator of island accretion is the freeboard, defined here as the difference between the island crest height and the extreme beach toe water level ($\eta_{2\%}$). Below a 0.5-m freeboard, crest lowering generally occurs, while freeboards between 0.5 and 0.8 m predominantly show accretion. Both crest accretion and island accretion are necessary to prevent the drowning of islands from SLR, although the challenges posed to island communities by the required flooding events are significant. Ignoring the morphological adaptation of the island leads to a threefold increase in the predicted inundation hours for Fiyoaree. Thus, for islands that can morphodynamically adjust to SLR, the assumption of a static island used in most climate risk studies results in a conservative estimate of future flood risk and adaptation timescales. However, given the computational expense of running advanced morphodynamic models, it is necessary to either to accelerate these models or to parametrize the island adjustment to sea level rise if they are to be included in large-scale flood risk assessments.

This study presents a methodological approach for long-term process-based modelling to gain further insight into the future coastal flood risk of coral islands. The study is limited by both hydrodynamic and morphodynamic model assumptions and schematizations. One-dimensional reef models, for instance, do not account for two-dimensional effects such as (artificial reef-induced) horizontal circulation cells and longshore currents, which can play a critical role in balancing wave-induced setup with offshore flow (e.g., Lowe *et al.*, 2010). Additionally, wave interactions resulting in energy transfer to low-frequency waves are overestimated in a 1D approach (e.g., Herbers *et al.*, 1994). However, these limitations do not detract from the primary focus of this paper which addresses the relative comparison of flood risk, rather than the absolute representation of flood risk, for different adaption strategies. Morphodynamic model limitations are related to the computation of sediment transport with the XBeach-gravel module, which is highly dependent on the hydraulic conductivity and Nielsen parameter, and these have been poorly validated for smaller sediment sizes (Masselink *et al.*, 2020). Additional experiments should inform on the accuracy of the morphological response of coral islands.

In terms of model schematization, there are limitations related to the representation of artificial reefs and the berm top barrier, as well as in the schematization of the wave climate. Artificial reefs were modelled as impermeable bed level elevations with enhanced roughness, following Roelvink *et al.* (2021). This approach does not account for flow

557 through the structure, leading to an overestimation of the wave height reduction, the setup, and of the reflection at
558 the structure, both for incident waves reflected offshore and for waves reflected from the beach towards the structure
559 and back. The overestimation of the setup across the structure, combined with the general overestimation of setup in
560 a 1D model, leads to underperformance of the structures in the model compared to their performance in real-world
561 conditions during energetic wave events. Flume, wave basin, and field experiments can help to determine the most
562 effective approaches to modelling these structures with XBeach. The effectiveness of the berm top barrier is strongly
563 influenced by its initial design. The current design included a relatively steep island crest, which subsequently led to
564 substantial morphological adjustments when overwash events occurred compared to the wider, unrestored island
565 crest. Finally, a single realization of wave event sequences was made. The final island profile and flood frequency
566 could be significantly influenced by the specific sequence of wave events, not explored here because of
567 computational limitations.

568
569 The ability of coral islands to adapt to SLR is strongly influenced by the rate and total amount of SLR. The
570 modelled SLR of 0.5 m by 2100, assuming a constant SLR rate of 6.25 mm/yr, represents the median prediction for
571 an intermediate emission scenario (SSP2-4.5). At higher rates of SLR (3.7 mm/yr SLR has already been observed
572 between 2006 and 2018; *IPCC, 2021*), islands may not be able to keep up with SLR, leading to a rapid rollover and
573 destruction of the island (*Masselink et al., 2020*). Furthermore, for higher sea levels, especially when combined with
574 limited reef flat accumulation due to coral degradation, the reef coast will be exposed to greatly increased wave
575 energy (e.g., *Merrifield et al., 2014*), rendering the islands highly vulnerable to extreme events. The impact of SLR
576 rates on coral island morphology is likely to be influenced by adaptation measures. Artificial reefs may provide
577 better protection at higher SLR rates compared to fore reef restoration, as they are most efficient at attenuating wave
578 energy, especially for larger water depths. Therefore, they require less morphological updating of the island to
579 achieve its equilibrium crest height, minimizing the risk of extreme overwash conditions. These considerations
580 highlight the vulnerability of coral islands and emphasize the need to consider different rates of SLR, vertical reef
581 growth scenarios, and adaptation measures to fully assess the potential impacts of sea-level rise on these ecosystems
582 and coastal communities.

583
584 Coral islands rely on sufficient sediment supply to maintain their footprint. However, future sediment supply,
585 governed by the complex interplay of ecological and physical processes (*Kench et al., 2009; Tuck et al., 2021*), is
586 highly uncertain and depends on, among other things, storm climate, oceanic conditions, and the activity of bio-
587 eroding (e.g., parrotfish; *Perry et al., 2015*) and bio-producing (e.g., foraminifera; *Dawson et al., 2014*) organisms.
588 In addition, sediment supply from the reef system to the island can vary considerably over time and varies between
589 islands. For example, following storm or coral bleaching events, large pulses of sediment may be transported
590 towards the island, possibly followed by periods of low sediment supply. In the current model setup, no sediment
591 source is introduced to the reef flat, thus ignoring changes in island sediment volume and potentially overestimating
592 island retreat.

593
594 The morphological adjustment of coral islands to sea level rise through frequent flooding poses significant
595 ecological challenges and impacts. Flooding can have severe ecological impacts, including the disruption of
596 freshwater availability on the island (*Storlazzi et al., 2018*) and the destruction of agricultural activities (e.g., *Zahir*
597 *et al., 2006*). Effective reef management is essential to preserve and restore crucial ecosystem services provided by
598 coral reefs, such as protection from storms and erosion, food provision, and income generation. In addition to
599 reducing carbon emissions, this entails adopting local measures to manage solid waste disposal, minimizing nutrient
600 and sediment-rich runoff, controlling siltation from dredging activities, and eliminating destructive fishing practices
601 (e.g., *Mcleod et al., 2019*). For restoration, it is important to select suitable coral reef species that can withstand
602 elevated temperatures and increased acidity. Examples of this include porites, a coral species found on the shallow
603 reef flat and in the warm, shallow inter-island channels of Fioyaree.

604
605 To ensure the continued habitation of atoll islands that naturally adjust to sea level rise through overwash-induced
606 accretion, flexibility and adaptability are required of both the island infrastructure and its communities. Flexible
607 infrastructure can be achieved through a variety of measures, such as building houses on stilts (*Esteban et al., 2019*),
608 flood protection at the household level (raising doorstep levels, implementing local flood walls; *Brown et al., 2020*),
609 and nature-based coastal protection measures (*Barnett et al., 2022*). Social adaptation is also a key factor in allowing
610 for natural island dynamics. An integral part of this adaptation is the implementation of early warning systems, to
611 enable island inhabitants to prepare effectively for potential flooding events (*Winter et al., 2020; Hoeke et al.,*
612 *2021*). Allowing for natural island dynamics seems feasible in the context of the Maldives (except for the

613 northernmost atolls), with its relatively consistent wave climate and lack of catastrophic cyclones (*Wadey et al.,*
 614 *2017*). Other island communities may face different challenges, particularly in parts of the Pacific where cyclones
 615 can destroy entire islands (*Ford et al., 2014*). Nevertheless, if embraced and implemented by local communities,
 616 allowing for natural island dynamics could potentially extend the habitability of islands in the Maldives beyond
 617 current projections.

618 **Acknowledgments**

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 620 input files and raw output files) in the study are publicly available at the Zenodo repository via
 621 <https://doi.org/10.5281/zenodo.10495462> (*Roelvink, 2024*). The XBeach software is freely available at
 622 download.deltares.nl/xbeach.

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